



BAHIA STATE UNIVERSITY
DEPARTMENT OF EXACT AND EARTH SCIENCES II
GRADUATE PROGRAM IN MODELING AND SIMULATION OF BIOSYSTEMS



KAROLINE DE MENEZES REBELLO

Modeling the Potential Distribution of *Moquiniastrum*
oligocephalum (Gardner) G. Sancho (Asteraceae: Gochnatieae) in Brazil

Alagoinhas, Bahia

2024

KAROLINE DE MENEZES REBELLO

Modeling the Potential Distribution of *Moquiniastrum*
oligocephalum (Gardner) G. Sancho (Asteraceae: Gochnatieae) in Brazil

Master's dissertation presented to the State University of Bahia, Course in Modeling and Simulation of Biosystems, for the degree of Master in Modeling and Simulation of Biosystems.

Field of Knowledge: Interdisciplinary; Conservation Biology.

Line of Research: Biosystems Analysis.

Supervisor: Dr. Gracineide Selma Santos de Almeida

Co-Orientator: Dr. Gustavo Reis de Brito

Alagoinhas, Bahia

2024

Sistema de Bibliotecas da UNEB
Biblioteca Carlos Drummond de Andrade – *Campus II*
Manoela Ribeiro Vieira
Bibliotecária - CRB 5/1768

R292m Rebello, Karoline de Menezes
Modelagem de Distribuição Potencial de *Moquiniastrum oligocephalum* no Brasil /
Karoline de Menezes Rebello – Alagoinhas, 2024
80f.:il

Orientadora: Profª Drª Gracineide Selma Santos de Almeida.
Coorientador: Profª Drª Gustavo Reis de Brito

Dissertação (Mestrado) – Universidade do Estado da Bahia, Departamento de Ciências Exatas e da Terra. Programa de Pós-Graduação em Modelagem e Simulação de Biosistemas. Mestrado em Modelagem e Simulação de Biosistemas – Alagoinhas, 2024.

I. Nicho ecológico – Modelagem 2. Conservação 3. Adequabilidade Ambiental I. Almeida, Gracineide Selma Santos de. II. Brito, Gustavo Reis de III. Universidade do Estado da Bahia – Departamento de Ciências Exatas e da Terra – Campus II. IV. TÍTULO

CDD – 574.5

FOLHA DE APROVAÇÃO
"ANÁLISE DA DISTRIBUIÇÃO DE MOQUINIASTRUM OLIGOCEPHALUM GARDNER G. SANCHO (ASTERACEAE: GOCHNATIEAE) NO BRASIL POR MEIO DA MODELAGEM DE NICHOS ECOLÓGICO"

KAROLINE DE MENEZES REBELLO

Dissertação apresentada ao Programa de Pós-Graduação em Modelagem e Simulação de Biosistemas – PPGMSB, em 28 de fevereiro de 2024, como requisito parcial para obtenção do grau de Mestre em Modelagem e Simulação de Biosistemas pela Universidade do Estado da Bahia, conforme avaliação da Banca Examinadora:

⋮

Professor(a) Dr.(a) GRACINEIDE SELMA SANTOS DE ALMEIDA
UNEB
Doutorado em Botânica
Universidade Federal de Viçosa

Professor Dr. GUSTAVO REIS DE BRITO
UNESP
Doutorado em Biociências
Universidade Estadual Paulista Júlio de Mesquita Filho

Professor(a) Dr.(a) ELTAMARA SOUZA DA CONCEIÇÃO
UNEB
Doutorado em Entomologia
Universidade Federal de Viçosa

Professor(a) Dr.(a) MARLETE MOREIRA MENDES IVANOV
UFPI
Doutorado em Ecologia e Recursos Naturais
Universidade Federal do Piauí

ABSTRACT

Ecological Niche Modeling (ENM) encompasses mathematical and computational processes guided by algorithms that utilize species occurrence data and environmental variables to establish correlations between distribution and environmental factors. This approach has fundamental implications for the study of biodiversity, decision-making in conservation, and understanding the ecological requirements of species. The application of ENM can aid in comprehending the distribution patterns of species within a biome or ecosystem through regional analysis, and it holds potential for application to other poorly studied tropical species, particularly those under the pressure of logging activities, such as *Moquiniastrum oligocephalum* (Gardner) G. Sancho (Asteraceae: Gochnatieae), commonly known as candeia, in Northeast Brazil. This species typically exhibits an arboreal habit, is ginodioecious, and comprises capitula arranged in reduced axillary panicles, with a biseriate pappus. The objective of this study was to investigate the distribution of *Moquiniastrum oligocephalum* in Brazil by analyzing the environmental suitability for the occurrence of the species, calculated through ENM. The methodology employed for generating environmental suitability models and analyzing the geographical distribution of *Moquiniastrum oligocephalum* proved satisfactory. The joint analysis focused on climatic conditions (Worldclim), combined with georeferenced occurrence points obtained from databases such as GBIF and speciesLink. The statistical validation of the models was based on the metrics of Area Under the Curve (AUC) and True Skill Statistic (TSS), yielding values of 0.97 and 0.87, respectively. The contribution values of the variables to the models were calculated through Relative Variable Importance, with the most significant variables identified as BIO1 - annual mean temperature (46.5%) and BIO4 - temperature seasonality (44.9%). From the generated models, it was concluded that there is greater suitability (exceeding 50%) for environments in the Caatinga and Atlantic Forest of Northeast Brazil, particularly in the states of Bahia, Sergipe, and Pernambuco. Thus, we can consider the Northeast region as a center for the dispersal of the species, especially in the state of Bahia, where significant environmental suitability was observed in the northern portion of the eastern coast and low environmental suitability in the southern and western areas of the state. For other regions of Brazil, results indicate reduced environmental suitability compared to the Northeast. No records of the species were found in the occurrence databases for the Central-West and Southern regions of Brazil, and there are few records for the Southeastern region. This study has thus expanded knowledge regarding the distribution of the species, providing essential support for conservation efforts in Brazil.

Keywords: Conservation; Ecological Niche Modeling; Environmental Suitability; MaxEnt.

RESUMO

A Modelagem de Nicho Ecológico (MNE), que consiste em processos matemáticos e computacionais guiados por algoritmos que utilizam dados de ocorrência das espécies e variáveis ambientais, para traçar correlações entre distribuição e ambiente. Possui implicações fundamentais para o estudo da biodiversidade, tomada de decisão em conservação e para a compreensão dos requerimentos ecológicos das espécies. O uso de MNE pode auxiliar no entendimento dos padrões de distribuição das espécies de um bioma ou ecossistema em uma análise regional, além de possuir potencial de aplicação para outras espécies tropicais ainda pouco estudadas, sobretudo aquelas que estão sob pressão da atividade madeireira, a exemplo da espécie *Moquiniastrum oligocephalum* (Gardner) G. Sancho (Asteraceae: Gochnatieae), espécie popularmente conhecida como candeia, no Nordeste brasileiro. Esta apresenta-se normalmente com hábito arbóreo, ginodióica, composta por capitulescências agrupadas em panículas reduzidas, axilares e pápus bisseriado. Este estudo teve como objetivo investigar a distribuição de *M. oligocephalum* para o Brasil, a partir da análise da adequabilidade ambiental para ocorrência da espécie, calculado a partir da MNE. A metodologia empregada para a geração de modelos de adequabilidade ambiental e análise da distribuição geográfica de *M. oligocephalum* foi satisfatória. A análise conjunta pautou-se no recorte de condições climáticas (Worldclim), combinado a pontos de ocorrência georreferenciadas, obtidos nos bancos de dados do GBIF e speciesLink. A validação estatística dos modelos foi baseada nas métricas da *Area Under the Curve* (AUC) e *True Skill Statistic* (TSS), com valores de 0,97 e 0,87, respectivamente. Os valores de contribuição das variáveis para os modelos foram calculados por meio da Importância Relativa das Variáveis, sendo consideradas as mais significativas: BIO1 - temperatura média anual (46.5%) e BIO4 - sazonalidade da temperatura (44.9%). A partir dos modelos gerados foi possível concluir que há maior adequabilidade (superior a 50%) para ambientes de Caatinga e Mata Atlântica nordestina no Brasil, principalmente nos estados de Bahia, Sergipe e Pernambuco. Podemos assim, considerar a região Nordeste como centro de dispersão da espécie, especialmente o estado da Bahia onde foi observada significativa adequabilidade ambiental na porção norte da costa leste do estado e baixa adequabilidade ambiental em áreas ao Sul e Oeste baiano. Para demais regiões brasileiras, resultados mostram reduzida adequabilidade ambiental, em relação ao Nordeste. Não foram encontrados registros nos bancos de dados da ocorrência da espécie nas regiões Centro Oeste e Sul do Brasil e há poucos registros da espécie para a região Sudeste. Foi possível, a partir desse estudo, ampliar o conhecimento acerca da distribuição da espécie para fornecer subsídios para as ações de conservação no Brasil.

Palavras-chave: Conservação; Modelagem de Nicho Ecológico; Adequabilidade Ambiental; MaxEnt.

I dedicate it to my uncle José Sampaio Rebello, my grandmother Alzira Menezes, and my dog friend Bucão, who left this world while I was looking for strength to write this scientific research.

ACKNOWLEDGMENTS

Firstly, I thank the spirituality that guides me on this path of many achievements, renunciations, and learnings, making my life an eternal gift.

I extend my gratitude to my advisor, Prof. Dr. Gracineide Selma, and Prof. Dr. Gustavo Reis for their excellent work, guidance on this research, encouragement, and patience. Time flew by, and they showed me how to overcome challenges and become a better person. I am also grateful for the existence of the plant species *Moquiniastrum oligocephalum*, which teaches me the importance of resilience. My gratitude extends to everyone who contributed to the Teaching Internship in the Ethnobotany discipline during my Master's program.

I am deeply grateful to my beloved family, especially my sisters Carine Rebello and Aline Rebello. To my new friends in postgraduate studies and at UNEB, particularly Amanda de Jesus and my "Cristinas." To my dear Dona Marina, from Alagoinhas, Bahia, for welcoming me into her home many times, for our conversations, and for the lessons in Ethnobotany.

I acknowledge CAPES for the financial support and commitment.

I also thank the healthcare professionals who accompanied me through difficult times, especially Madalena Ribeiro and Dr. Yuri Copelasso.

To my great friend and artist Eric Assmar from Salvador, Bahia, for always holding my hand in many moments of life. To the beloved musicians of my band, for understanding my absences. To my dear friend, Professor Paulo Roberto, for his encouragement and for providing various grammatical corrections.

A special thank you to Professors Dr. Eltamara Conceição and Dr. Marlete Moreira for their encouragement. Gratitude to Prof. Dr. Iramaia de Santana and Prof. Dr. José Fontoura for their teachings and dedication in supporting me.

I am thankful to the Postgraduate Program in Modeling and Simulation of Biosystems for the classes, various training sessions, and guidance, and for the opportunity and support in participating in the 2023 National Congress of Botany.

I am confident that all the dedication invested here was worthwhile.

SUMMARY

ABSTRACT

RESUMO

1. INTRODUCTION	1
2. OBJECTIVES	4
2.1 General	
2.2 Specific	
3. THEORETICAL BACKGROUND	
3.1. The Asteraceae family in Brazil and worldwide	5
3.2 Species distribution patterns	9
3.3 Ecological Niche Modeling (ENM)	11
4. METHODOLOGY	
4.1 Study area	15
4.2 Survey of species occurrence data	16
4.3 Environmental variables	18
4.4 Ecological Niche Modeling (ENM)	20
5. RESULTS AND DISCUSSION	
5.1 Geographical distribution of <i>Moquiniastrum oligocephalum</i> in Brazil	22
5.2 Ecological Niche Modeling of <i>Moquiniastrum oligocephalum</i> in Brazil	31
5.3 Priority areas for conservation of the species in Brazil	37
6. FINAL CONSIDERATIONS	42
7. REFERENCES	44

1. INTRODUCTION

The knowledge about the geographical distribution of species is fundamental for understanding the ecological and evolutionary determinants of biodiversity's spatial patterns and is essential for biological conservation (Ferreira, 2009). The complexity of biological systems follows a series of parameters that enable the survival and development of living beings. For example, abiotic factors such as temperature, humidity, and altitude, together with biotic factors like migration, competition, and dispersal, directly influence the distribution of organisms.

Habitat fragmentation, land-use changes, and climate change threaten the existence and perpetuation of species worldwide. With the intensification of global warming, it is notable that various biomes around the Earth are undergoing transformations, indicating changes that may increasingly lead to desertification processes and the occurrence of extreme events such as severe droughts and intense rainfall (Lacerda *et al.*, 2016). Conversely, the growing need for ecosystem protection and restoration demands new technologies, which are being improved and can relate environmental characteristics to species occurrences (Giannini *et al.*, 2012; Moreno-Fernández *et al.*, 2016), as environmental changes have especially endangered the distribution of plant species.

Knowledge about species distribution, particularly plants, remains scarce in Brazil (Diniz-Filho *et al.*, 2013; Hortal *et al.*, 2014). One of the challenges faced is the lack of knowledge about the richness and geographical distribution of a considerable number of species. However, there have been significant advances towards a better understanding of phylogenetic relationships within the Asteraceae family.

Moquiniastrum oligocephalum (Gardner) G. Sancho is a species in the Asteraceae family, of economic importance, commonly known in Northeast Brazil as "candeia" or "candeeiro." It typically presents tree-like characteristics, is gynodioecious, and comprises capitulescences grouped in reduced, axillary panicles, shorter than the leaves, rarely bracteate, containing tubular cream-colored flowers and biseriate pappus (Flora e Funga do Brasil, 2023). This species is sought after for its timber potential, highlighting the need for greater efforts in environmental protection to avoid the impacts of population reduction in its niche.

Thus, it is important to invest in and reinforce measures that enable biodiversity conservation, combining efforts to maintain priority areas (Peterson et al., 2011), by identifying likely species occurrence sites and planning management and conservation actions (Funk and Richardson, 2002).

A suitable measure and useful strategy to fill gaps in the knowledge produced about plant species distribution is Ecological Niche Modeling (ENM). ENM consists of a set of modern techniques that use data from various knowledge areas, resulting from the integration of biological knowledge and various technologies (Giannini et al., 2011), to obtain models based on mathematical approximations of the ecological niche (Sillero et al., 2021), allowing discussions about the probabilistic distribution of one or more species in the environment.

The use and application of ecological models related to species distribution have increased. With the growth of scientific research in ecology, computational development, and the use of methods based on geographic information systems, species distribution modeling has gained prominence, mainly due to the availability of algorithms for data analysis, also aiding in understanding species' ecological requirements. This results in fundamental implications for biodiversity studies and determining priority conservation areas (Ferreira, 2009; Anderson, 2013).

The construction of ENM requires the use of environmental variables in conjunction with georeferenced information (Anderson et al., 2003), as well as the use of algorithms that estimate the relationship between this data and map habitat suitability to represent required conditions (Franklin and Miller, 2010). It is a very useful tool for addressing applied ecology, biogeography, and conservation questions due to its ecological and evolutionary foundations (Guisan and Thuiller, 2005). Thus, it provides well-founded responses to current challenges of environmental threats that many species face (Giannini, 2012), and is important in biological analyses, particularly in supporting conservation and management interventions, as it informs about environmental requirements and species distribution, demonstrated in mapping, which are crucial aspects in biological analyses and support for conservation and management interventions. Additionally, knowing plant species and their extinction threat classification is essential for informing conservation policies.

In Brazil, one form of flora protection is the establishment of Conservation Units (UC), aimed at contributing to species conservation and protecting biodiversity (Brasil, 2000). The identification of priority conservation areas in Brazil aims to recognize those locations or regions with unique natural attributes, considered critical for maintaining regional biodiversity (Brasil, 2004). From the development of a technical management plan document, for example, it is possible to establish zoning and guide the actions for the use of UC natural resources (Brasil, 2000). However, selecting priority conservation areas depends on the knowledge of the biodiversity to be conserved.

Considering this information, this unprecedented study was developed to present data on the updated distribution of the species in Brazilian territory. The promising results can be used to guide conservation actions and maintain priority areas in Brazil, especially in the state of Bahia, which has the highest number of species occurrence records.

2. OBJECTIVES

2.1 General

To determine the geographical distribution of *Moquiniastrum oligocephalum* (Gardner) G. Sancho in Brazil, based on occurrence records and the application of the Ecological Niche Modeling (ENM) technique.

2.2 Specific

- ✓ To analyze the geographical distribution of the species and the bioclimatic variables that determine its occurrence in Brazil;
- ✓ Draw up a geographical distribution map of the species in Brazil;
- ✓ Generate Environmental Suitability models based on Ecological Niche Modeling to analyze the geographical distribution of the species in Brazil;
- ✓ Investigate and locate possible priority areas for conservation of the species in Brazil.

3. THEORETICAL BACKGROUND

3.1. The Asteraceae Family in Brazil and Worldwide

Asteraceae Bercht. & J. Presl is one of the most diverse families of Angiosperms (BFG, 2018). It constitutes a monophyletic group, composed of approximately 25,000-30,000 species, grouped into about 1,700 genera (Funk et al., 2009). In Brazil, the family is grouped into 327 genera and about 2,208 species (Flora e Funga do Brasil, 2023). Bremer (1994) divided the family into three subfamilies and 17 tribes. Following the publication of Funk et al. (2009), a classification of Asteraceae was recognized, consisting of 13 subfamilies and 44 tribes, representing about 10% of the world's vascular plants (Funk et al., 2009; Panero et al., 2014).

Asteraceae is characterized by its capitulum inflorescence, with flowers arranged on a receptacle surrounded by bracts. The capitula are composed of sessile flowers, solitary at the apex of the floral peduncle (scape), or usually by few to many capitula in corymbiform, spicate, paniculate, racemose, umbelliform cymose capitulescences (Flora e Funga do Brasil, 2023). The anthers are syngenesious, with secondary pollen exposure, and the ovary is bicarpellate and inferior, with a single ovule of basal placentation that develops into an achene, generally with a pappus (Funk et al., 2005, 2009; Roque and Bautista, 2008).

It has a wide geographical distribution, being present on all continents except Antarctica (Panero, 2008; Roque and Bautista, 2008). It is most abundant in temperate and semi-arid regions of the tropics and subtropics, recorded in open vegetation areas such as fields, high-altitude regions, and low-altitude tropical rainforests (Funk et al., 2009b; GBIF, 2023). However, despite the considerable level of information about Brazilian biodiversity, gaps remain that need to be filled, including those to support conservation actions and improve the understanding of phylogenetic relationships in Asteraceae (Funk et al., 2009).

The efficiency in dispersion confers extreme importance to Asteraceae in understanding the recovery of degraded areas, where they act as pioneers in colonizing degraded environments and occur in clearings and forest edges (Heiden et al., 2007). Distributed in all biomes, the family occurs in Brazil in vegetation formations of Cerrado, Campos, Atlantic Forest, Restinga, Caatinga, swamp areas, riparian forests, and high-altitude forests (Moura and Roque, 2016). Second only to the Cerrado, among the country's phytogeographic domains, the Atlantic Forest is the second richest in Asteraceae, with 187 genera and 961 species (Roque et al., 2020).

Asteraceae exhibits herbaceous, subshrub, shrub, or tree habits, generally terrestrial, rarely epiphytic or aquatic, sometimes succulent. Asteraceae species are extensively studied

from botanical, chemical, and pharmacological perspectives (Bremer, 1994; Funk et al., 2009). Research reports the presence of some cytotoxic and anti-inflammatory activities (Funk *et al.*, 2009).

The family stands out worldwide for its economic, ornamental, and medicinal importance. Due to the high chemical diversity present and the great bioactive potential, one of the main uses of Asteraceae species is their employment as herbal medicine in popular medicine (Funk *et al.*, 2009). Additionally, many plants are edible, and the consumption and cultivation of leaves, stems, seeds for vegetable oil extraction are known, such as *Helianthus annuus* L. (sunflower); *Cynara scolymus* L. (artichoke); and *Lactuca sativa* L. (lettuce). Other species are considered medicinal plants, like the sweetener extracted from *Stevia rebaudiana*; ointments (*Calendula officinalis* L., *Arnica montana* L.); teas, such as chamomile (*Matricaria recutita* L.) (Panero and Funk, 2008; Moura and Roque, 2014). Phytochemical studies are known among *Moquiniastrum* species, such as *Moquiniastrum argentinum* (Cabrera) G. Sancho, *Moquiniastrum barrosoae* (Cabrera) G. Sancho, *Moquiniastrum blanchetianum* (DC.) G. Sancho, *Moquiniastrum haumanianum* (Cabrera) G. Sancho, *Moquiniastrum paniculatum* (Menos.) G. Sancho, *Moquiniastrum polymorphum* (Menos.) G. Sancho, *Moquiniastrum polymorphum* subsp. *floccosum* (Menos.) G. Sancho, *Moquiniastrum pulchrum* (Cabrera) G. Sancho, from which dozens of special metabolites have been described (Tamayose, 2019).

The production of substances from compound extraction has been reported for the Asteraceae family (Calabria et al., 2007), such as special metabolite classes like flavonoids and terpenoids, with the most common subclasses in Asteraceae being flavones and flavonols, which have been related to the evolutionary success of its individuals (Funk et al., 2009; Tamayose, 2019). Flavonoids generally have great ecological importance for plants, acting as attractants for pollinators, seed and fruit dispersers, signaling in plant-plant and plant-microorganism interactions, and providing UV radiation protection (Funk et al., 2009).

According to Tamayose (2019), experimental procedures with the species *M. oligocephalum* allowed the identification of 20 substances, extracted from samples collected in the Caatinga phytogeographical domain region, in the municipality of Morro do Chapéu, Bahia. This was the first report of the study of epicuticular waxes in *Moquiniastrum* species, especially in *M. oligocephalum*.

In recent years, several classifications have been proposed for the species of the family. Asteraceae constitutes an easily recognizable and clearly monophyletic group (Judd et al., 2009). Recent molecular studies have proposed arrangements for its classification (Moura and Roque, 2014). Thus, the advent of molecular biology brought new changes to the circumscription of

Asteraceae genera, tribes, and subfamilies, now based on genetic data (Panero and Funk, 2008; Funk et al., 2009).

Despite significant advances in understanding the current phylogenetic relationships in Asteraceae, one of the major problems faced in Brazil is the lack of knowledge about the geographical distribution of a large part of the species and lack of knowledge (inventories) about the richness of genera and species. Additionally, there is high morphological variation and complexity (Roque et al., 2016, 2017). Information gaps hinder the application of conservation actions, and the conservation of the species of interest is limited to areas where it has already been recorded (Moscoso, 2012).

The subfamily Gochnatioideae currently has 103 species and is considered one of the most basal in Asteraceae and a sister group to approximately 96% of the family's species (Panero and Funk, 2008; Sancho and Freire, 2009; Funk et al., 2014). Many morphological studies have not resolved the situation among the sections of the genus *Gochnatia*, aimed at establishing generic and infrageneric relationships (Funk et al., 2014). Freire, Katinas, and Sancho (2002) published a morphological analysis of *Gochnatia* species defined by Cabrera (1971), called the "*Gochnatia* complex."

The genera of Gochnatioideae have characteristics such as an apiculate connective appendage, rounded styler branches, and dorsally glabrous, with the tribe's apomorphy based on these two characters (Funk et al., 2014). According to the same author, based on phylogenetic studies using molecular and morphological evidence, *Gochnatia* was reduced to approximately 40 species.

In the 1970s, a revision of the Gochnatinae tribe species led to the redefinition of the *Gochnatia* genus, grouping 70 species into six sections, one of which was named *Moquiniastrum*. Recent molecular phylogeny studies showed that the species of the *Moquiniastrum* section were very close to each other and different from the other species of the genus, determining its elevation to the genus level (Cabrera and Klein, 1973; Sancho, Funk, and Roque, 2013).

Initially, the tribe consisted of four genera: *Cnicothamnus* Griseb. (1874), *Cyclolepis* Gilles ex D. Don (1832), *Gochnatia* Kunth (1818), and *Richterago* Kuntze (1891). Currently, the genera included in Gochnatioideae vary in their morphology and habitat: *Gochnatia*, *Pentaphorus*, *Anastraphia*, *Moquiniastrum*, *Richterago*, *Cnicothamnus* (sister group to the *Moquiniastrum* + *Richterago* clade), and *Cyclolepis*. In the phylogeny of the Gochnatioideae subfamily, the *Moquiniastrum* and *Richterago* genera are considered sister groups with statistical support (Funk et al., 2014).

A phylogenetic study of Gochnatieae by Funk et al. (2014) included more than 60% of the tribe's species (represented by 112 samples) and sampled six molecular markers for comparative phylogenetics. This effort confirmed the monophyly and phylogenetic position of Gochnatieae, which currently has seven genera, including the *Moquiniastrum* (Cabrera) G. Sancho genus, and about 80 species, with geographical distribution restricted to the American continent (Funk et al., 2014; Roque et al., 2016; Flora e Funga do Brasil 2024).

In the current classification, the genus is monophyletic and presents differences from Gochnatia, such as gynodioecy; indumentum with 2-5-branched trichomes and usually paniculate capitulescence, actinomorphic corolla deeply lobed, 2–3-seriate pappus, and usually paniculate synflorescence (Sancho, Funk, and Roque, 2013; Freitas, 2014). The sexual complexity of *Moquiniastrum* is unique, with its species essentially classified into five groups according to the sexuality of the flowers: gynomonoeocious, gynodioecious, homogamous, hermaphroditic, polygamous (Sancho, 2000). According to Gisele Sancho, the genus is the only one that has functionally female flowers due to the presence of rudimentary anthers (staminodes), considered a model for studying the evolution of the plant sexual system.

The ancestral distribution of *Moquiniastrum* (Cabrera) G. Sancho corresponds to a large area covering eastern South America and the current central Andes, about 32 million years ago (Gostel et al., 2022). The genus is not endemic to Brazil, also occurring in Argentina, Bolivia, Peru, Venezuela, and Uruguay. In its worldwide distribution, *Moquiniastrum* consists of 22 species, with 20 species occurring in eastern Brazil (Flora do Brasil, 2024), including the new species *Moquiniastrum glabrum* Roque, Neves & A. Teles (Roque et al., 2019).

In Brazil, the genera *Moquiniastrum* and *Richterago* occur (Funk et al., 2014). The publication by Sancho, Funk, and Roque (2013) provided a nomenclatural rearrangement by segregating *Moquiniastrum* and elevating it to the category of a genus (Freitas, 2014). It is distributed in the phytogeographic domains of the Cerrado (Minas Gerais, Bahia, Goiás, Paraná, São Paulo), Pampa (Rio Grande do Sul), and also in regions of the Atlantic Forest (from the Northeast to the South of the country) (Sancho et al., 2013, 2014; Flora do Brasil, 2023).

Of the 20 species of the genus *Moquiniastrum* that occur in Brazil, nine are found in the state of Bahia: *Moquiniastrum barrosoae* (Cabrera) G. Sancho, *Moquiniastrum blanchetianum* (DC.) G. Sancho, *Moquiniastrum densicephalum* (Cabrera) G. Sancho, *Moquiniastrum floribundum* (Cabrera) G. Sancho, *Moquiniastrum paniculatum* (Less) G. Sancho, *Moquiniastrum polymorphum* (Less.) G. Sancho, *Moquiniastrum discolor* (Baker) G. Sancho, *Moquiniastrum glabrum* Roque, Neves & A. Teles, including *Moquiniastrum oligocephalum* (Gardner) G. Sancho (Sancho et al., 2013, 2014; Roque, 2019; *Specieslink*, 2024).

Moquiniastrum oligocephalum (Gardner) G. Sancho, popularly known in northeastern Brazil as "candeia," is morphologically well-defined within the homogeneous group to which it belongs. The species is distinguished by its typically arboreal habit, tomentose branches, discolorous leaves with branched trichomes, paniculate capitulescence, tubular flowers, and for being gynodioecious (Bremer, 1994; Sancho, 2013; Freitas, 2014; Roque et al., 2019; Flora do Brasil, 2023). According to Sancho, Funk, and Roque (2013), the species has important dispersion centers in Brazil, and the genus also occurs in the following countries according to Flora e Funga do Brasil (2023): Argentina, Uruguay, Bolivia, Paraguay, Peru, Venezuela.

M. oligocephalum has been increasingly sought after for its timber potential, and due to recurring plant extraction, its population tends to be further reduced, especially in Atlantic Forest areas in northeastern Brazil, where local populations are currently threatened with extinction. In northeastern Brazil, this phytodomain presents various physiognomies and according to Tabarelli et al. (2006), more than 46% of the mapped Atlantic Forest remnants are located in the state of Bahia. However, more work needs to be done to increase sampling efforts to enable the recovery of areas and greater promotion of floristic conservation. Woody species, such as *M. oligocephalum*, are especially important in semi-arid regions for protection against desertification and resilience to climate change (FAO, 2016).

3.2. Patterns of Species Distribution

The distribution of organisms across the different environments of the Earth reflects part of the history of geological, ecological, and evolutionary events that have occurred together. Generally, it is known that the geographic distribution of species results from complex ecological relationships involving: soil type, microclimate, macroclimate, altitude, plate tectonics, riverbed narrowing/widening, glaciations, vegetation physiognomies, climate, humidity, ocean currents, salinity, precipitation and day length, food availability, as well as the dispersal capacity of organisms and the influence of geographic and climatic factors such as latitude, altitude, ocean currents, and air masses (Brown and Lemolino, 2006; Soberón, 2010). Additionally, dispersal capacity, species interactions (biotic factors), and anthropogenic impacts are influential (Morellato et al., 2000; Pearson, 2006; Soberón, 2010; Wisz et al., 2013; Wang et al., 2017).

Studies involving dispersal syndromes, for example, have been considered for understanding biogeographic patterns because they are directly related to the distribution of

species (Diniz-Filho et al., 2009). Ecologists and biogeographers have been striving over the years to understand the processes involved in the generation and maintenance of the current pattern of biodiversity distribution (Hutchinson, 1959; Diniz-Filho et al., 2009). Some authors emphasize that the distribution of most plant species is mainly controlled by climatic factors such as temperature and precipitation (Bañuelos et al., 2004; Kendal et al., 2012). Regarding neotropical forests, according to Gentry (1983), these environments show patterns in taxonomic composition, diversity, and species dispersal ecology.

According to Neves et al. (2015) and Queiroz et al. (2017), variations in precipitation and temperature can lead to significant changes in species composition and ecological strategies (Santiago et al., 2016). Antonelli (2017) highlights that climate influences species distribution patterns due to its variations, resulting in different scenarios in different periods within the same environment. Thus, biotic and abiotic relationships determinately influence species' geographic distribution (Morellato et al., 2000; Soberón, 2010), and naturally, there are temporal variations in speciation, extinction, and migration rates (Condamine et al., 2018).

Brazil has a great biological diversity (Mittermeier et al., 1998) and a wide variety of environmental factors that operate at different intensities and scales, adjusting the occurrence limits of organisms. The northeastern region of Brazil contains various morphoclimatic domains, rich in ecosystems (Ab'Sáber, 1971, 1977). The coastal forest establishes itself in a narrow strip along the Atlantic coast. This forest encompasses two main vegetation types: Dense Ombrophilous Forest (closer to the coast) and Semideciduous Forest, which forms a narrow belt of drier forest inland (Veloso et al., 1991).

The degree of endemism, according to Gentry (1986), is an important criterion used for selecting priority areas for conservation. Thus, diagnosing probable endemisms and information on the geographic distribution of plant species contribute to promoting identification, diagnosis, and possible recovery programs for areas considered conservation priorities. Diniz-Filho et al. (2009) points out that understanding diversity patterns on large scales and the processes involved in their origin and maintenance are important for establishing more efficient biodiversity conservation programs. Moscoso (2012) emphasizes that knowledge about the distribution of plant species is essential for biodiversity management and conservation planning. Funk and Richardson (2002) assert that understanding biodiversity patterns can be key to conserving the remaining existing species, especially in tropical areas.

There are still many species to be described and cataloged, and knowledge about species identity and distribution is still developing, with gaps in biodiversity study (Cardoso et al., 2011; Hortal et al., 2014). One strategy that can help fill these gaps is to estimate the geographic

distribution of species and characterize favorable environmental conditions for species occurrence, possibly locating and identifying how specific environments suitable for the survival of studied organisms are distributed in space (Pearson, 2007).

3.3. Ecological Niche Modeling (ENM)

Ecological niche modeling is defined as the inference of the geographic distribution of one or more organisms through methods or algorithms (Guisan and Zimmermann, 2000; Peterson and Soberón, 2012). According to Siqueira's research (2005), based on biological data (i.e., species occurrence points) and environmental data (i.e., thematic maps), algorithms can be applied in the processes of geographic distribution modeling with the aim of modeling the fundamental niche and estimating probable areas of species occurrence. To do this, they combine species occurrence data with the environmental variables of these locations, seeking to identify areas with suitable environmental conditions for population survival (Elith and Leathwick, 2009; Pearson, 2007; Thuiller, 2007).

Grinnell (1917) described the environment occupied by the bird *Toxostoma redivivum*, characterizing the niche as the environmental conditions that allowed its survival. In the case of the "Grinnellian" niche concept, interactions between living beings or the influence of environmental resources are not considered, as it focuses on the range of variables and the set of large-scale abiotic components that allow species to exist (Soberón, 2007). Based on Grinnell's niche concept (1917), the areas indicated in the modeling are habitats and satisfactory locations for species occurrence (Phillips, 2008).

Hutchinson (1957) defined the niche as a multidimensional space, within which there are several axes representing different environmental variables, both physical and biological, that allow the survival of a given species and limit its abundance and distribution, such as limiting factors like light, temperature, and food resources. Hutchinson's proposed concept considers the fundamental ecological niche, meaning the one that the species can occupy and encompasses all the conditions and resources that allow a particular species to exist and reproduce in the absence of other species.

Authors such as Soberón and Peterson (2005) mention that niche models provide an approximation to the species' fundamental niche, while others argue that modeling is the spatial representation of the realized niche (Guisan and Zimmermann, 2000; Pearson and Dawson, 2003). Soberón and Peterson (2005) added the dispersal capacity to the interpretation of models and created the Biotic, Abiotic, and Movements (BAM) diagram.

The projection of model predictions onto the geographic space results in environmental suitability maps (Sillero, 2011), which indicate on a continuous numerical scale the similarity of the mapped environments with those where the species under study is present within the domain of the bioclimatic predictor variables used in the modeling. Thus, ecological niche models consist of a simplified approximation of complex processes and assume that the environmental variables of the study area represent an environmental approximation of a particular species' niche (Phillips et al., 2006).

According to Guisan and Thuiller (2005), the interest in computing to study the influence of environmental variables on species distributions was mainly concentrated in the 1970s, 1980s, and 1990s. It offers rapid and well-founded responses to the threats species face (Sillero, 2011; Giannini, 2012), having been used in species distribution analyses, providing information on ecosystems and biomes (Lima-Barreto, 2015; Sobral-Souza, 2018), linked to biogeographic studies (Siqueira and Durigan, 2007); species conservation (Engler, Guisan, and Rechsteiner, 2004), and aiding in the determination of priority areas for conservation (Chen, 2009).

Species occurrence data have been made available by the scientific community in recent years (Trainor et al., 2014). These actions resulted from international initiatives that primarily aimed at standardizing, sharing, and providing primary data from biological collections, museums, and herbaria (Graham et al., 2004), among them the Global Biodiversity Information Facility (GBIF) database on taxonomic information, whose primary goal is to make biodiversity data available. In Brazil, the *speciesLink* network was created in 2002 with the general objective of making data on Brazilian biodiversity available.

Initially, in addition to obtaining species occurrence data and delineating the study area, considering environmental parameters and species biology for statistical analyses, it is necessary to select bioclimatic variables and one or more algorithms to trace potential correlations between the environment and distribution. Phillips (2008) discusses that statistically, the modeling result indicates whether a location is satisfactory for species occurrence and not exactly whether it is being occupied. Thus, the probability of species occurrence can be estimated.

Methodologies using computational algorithms are generally applied to represent primary data, species occurrence of interest, in distribution maps, thus indicating the probable presence or absence (Araújo and Guisan, 2006). These seek to establish non-random relationships between occurrence data and relevant environmental variables for the studied species, such as temperature, precipitation, and topography.

The statistical validation of models uses various techniques to assess the performance of algorithms and the predictive capacity of trained models. Among the methodologies presented in

the literature, the Area Under the Curve (AUC) and True Skill Statistics (TSS) metrics (Thuiller et al., 2009) are the most used in the statistical validation of ENMs and ENMs. The closer to 1.0 the AUC value, the better the model's performance (Phillips et al., 2006; Merow et al., 2013).

Regarding modeling procedures, important elements to ensure the quality of the final modeling include the proper definition of environmental layer resolution; accuracy of occurrence points; terrain and species characteristics (Elith and Leathwick, 2009); and the size of the geographic area to be analyzed (Chapman et al., 2005). According to Hernandez (2006), the influence of sample size demonstrates the need to better define a data set based on species biology knowledge and the algorithm's performance used to model it.

MaxEnt is an algorithm capable of estimating the probability of species occurrence, using an optimization procedure based on the principle of maximum entropy, thus relating the presence of species with environmental characteristics (Phillips, 2006) based on Shannon's entropy concept, following what was proposed by Phillips, Anderson, Schapire (2006), Steven et al. (2019), and Dai et al. (2022). This algorithm has a simple user interface and typically performs well, requiring only presence data (Wisniewski et al. 2008). It is one of the most popular tools for species distribution modeling (Merow et al., 2013) and has proven more effective than other programs in processing small data sets (Elith et al., 2006).

Environmental analysis techniques, using MaxEnt, can indicate environmental conditions similar to those where the species was found, represented on the environmental suitability map scale. Thus, the distribution of occurrences in the geographic space can be estimated from the environmental variables related to both the presence data and the landscape (Lima-Ribeiro and Diniz-Filho, 2012). The model generated for a given species is a continuous surface of values ranging from zero to 100, where high values indicate a greater probability of finding the species under study in the region (Guisan and Zimmermann, 2000). The algorithm execution considers that the species' distribution probability is uniform within the study area (background) (Elith et al., 2006).

According to Carvalho et al. (2019), the performance of each algorithm is related to both the type of species distribution modeled and previous studies by Araújo (2006); Elith et al. (2006) and Pearson et al. (2006). Therefore, the choice of algorithm should be based on the study question and the availability of occurrence data, with no consensus on the best algorithm to use, as the choice should be considered on a case-by-case basis (Ginannini et al., 2012).

There are several types of modeling algorithms, whose classification criteria lie in how they process the data. Due to the complexity of the process concerning reliable predictions, different algorithm and method approaches have been applied, such as bioclimatic envelope

algorithms like BIOCLIM (Busby, 1991); statistical fitting ones, like Generalized Linear Model – GLM (generalized linear models) (Wiley & Sons, 2006) and Generalized Additive Model – GAM (generalized additive models); machine learning ones (artificial neural networks) – ANN; support vector machine – SVM, decision trees – CART, random forest – RF and maximum entropy – MAXENT), as observed in the works of (Phillips et al., 2006; Paglia et al., 2012; Merow et al., 2014; Araújo, 2016; Cerdeira et al. 2018).

Understanding species distribution patterns is fundamental for conserving biological diversity (Ferreira, 2009). Thus, species distribution work using Ecological Niche Modeling is relevant for biological analyses, contributing to conservation and management actions, aiding decision-making aimed at biodiversity maintenance and rehabilitation in various contexts.

4. METHODOLOGY

4.1. Study Area

The Brazilian territory was the considered study area, with emphasis on the Northeast region, especially the state of Bahia, for which a modeling projection was carried out after analyzing the distribution data of the species *Moquiniastrium oligocephalum*. For the production of the species' geographic distribution maps, two locations were considered: the entire Brazilian territory and the state of Bahia, as this state has the highest number of occurrence records for the species compared to others. Thus, the geographic coordinates of the species' occurrence points were imported into the QGIS software (QGIS, 2023) and overlaid onto the geographic map of Brazil (IBGE, 2019) for better visualization and analysis of the species' spatial distribution in the Brazilian territory.

Considered the largest country in South America and in biodiversity globally, with a variety of terrestrial biomes (Amazon, Caatinga, Cerrado, Atlantic Forest, Pampas, and Pantanal) (Flora and Funga of Brazil, 2023), Brazil has a predominantly tropical climate (UNDP, 2020), and factors such as surface area, relief, and both latitudinal and longitudinal amplitude result in significant climatic variations throughout the Brazilian territory (Larousse, 2013). Rock outcrops are found in various climatic domains, mainly occurring in the Northeast region, eastern part of the country, both in the semi-arid and the Atlantic Forest domain (Araújo et al., 2008).

In the Brazilian Northeast, it is possible to find some very peculiar geographic mesoregions, whose climates range from the super humid coastal zones to the dry climate of the hinterlands in the region known as the Northeastern Semi-Arid (Zanella, 2014). The Brazilian Northeast is bordered to the North and East by the Atlantic Ocean. Due to its large extension and location (48° 05' W - 35° 02' W and 1° S - 18° 05' S), it is influenced by various atmospheric systems, among which the intertropical convergence zone, eastern waves, cold fronts, and sea breezes stand out (Roucou et al., 1996). The Atlantic Forest phytogeographic domain covers about 15% of the Brazilian territory and is distributed along the entire coast, from the north of the state of Piauí to the south of the state of Rio Grande do Sul, along with isolated areas in Goiás, Mato Grosso, and Minas Gerais (SOS Mata Atlântica, 2018). The Atlantic Forest has been severely impacted by human intervention, usually associated with the massive exploitation of its natural resources (Stehmann et al., 2009). The proximity to the coast provides the Atlantic Forest with constant humidity from the ocean and an important rainfall regime for the maintenance of a significant portion of Brazil's water sources.

The Caatinga phytogeographic domain is the only one genuinely from the Brazilian Northeast and occupies an area equivalent to 11% of the national territory (IBGE, 2019), extending over sedimentary bases, mountains, and plateaus (IBGE, 1985). The vegetation is characterized by seasonality and low rainfall indices (Sampaio, 1995; Rodal and Melo 1999), forming a mosaic of landscapes with high floristic diversity (Sampaio et al., 2002). As for the Cerrado phytogeographic domain, it can be characterized as Brazilian savanna vegetation and once represented about 23% of the national territory, second only to the Amazonian domain (Ratter et al., 1997). It occupies about 25% of the national territory (IBGE, 2019) and is present in all Brazilian regions. The Cerrado features landscapes composing a mosaic of phytophysionomies, determined by different soil types, climatic conditions, and the fire regime characteristics of each location (Ratter et al., 2003; Durigan et al., 2003a; Ribeiro and Walter, 2008), with only small areas of the Cerrado being legally protected (Bianchi and Haig, 2013).

4.2. Data Collection of *Moquiniastrum oligocephalum*

Moquiniastrum oligocephalum (Gardner) G. Sancho is a species of the Asteraceae family, of economic importance, known locally in the Brazilian Northeast as "candeia" or "candeeiro". The species is distinguished by its typically arboreal habit, tomentose branches, discoloured leaves with branched trichomes, paniculiform capitulescence, tubular flowers, and its gynodioecious nature (Roque et al., 2019; Flora do Brasil, 2023). According to Sancho, Funk, and Roque (2013), the species has important dispersal centers in Brazil, and the genus also occurs in the following countries: Argentina, Uruguay, Bolivia, Paraguay, Peru, and Venezuela.

The construction of an Ecological Niche Model requires the use of environmental variables along with georeferenced information regarding the species' presence in the study area (Soberón, 2010). Thus, in the first stage, proper filtering was carried out using occurrence databases, considering only georeferenced data (i.e., complete occurrences with latitude and longitude records).

Regarding the global and regional occurrence records data of *Moquiniastrum oligocephalum*, these were obtained from online databases such as *SpeciesLink* (www.specieslink.net) and the Global Biodiversity Information Facility - GBIF (www.gbif.org) in the form of geographic coordinates. The presence records were individually reviewed, and subsequently, duplicates and spatially correlated occurrences were removed to prepare a spreadsheet. The species' occurrences were processed using the tidyverse package (Wickham; RStudio, 2023) to remove duplicates, incorrect and/or missing coordinates, as well as

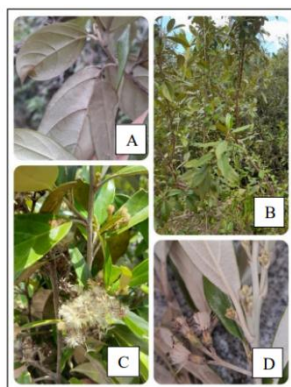
occurrences without location data. All procedures were performed in the R environment version 4.3.0 (R Core Team, 2023). Data related to the geographic distribution and Brazilian phytogeographic domains of *M. oligocephalum* were consulted on the Flora and Funga do Brasil website (<http://floradobrasil.jbrj.gov.br/reflora/>).

From the consultation of occurrence records of the study species in the databases, a total of 952 occurrence points were obtained. Initially, the occurrence matrix was filtered for the Brazilian territory, resulting in 482 selected points. Subsequently, data cleaning procedures were carried out to construct the definitive occurrence matrix for modeling, which consisted of 173 unique presence points for *M. oligocephalum*, structured in a comma-separated values (CSV) file.

The geographical thinning to remove sampling bias was performed using the spThin package (Aiello-Lammens et al., 2015), considering a radius of 10 km around each occurrence, with this process replicated 100 times by the best sample set selection algorithm.

The voucher material of *M. oligocephalum* was properly collected in the Atlantic Forest environment, within the vegetation complex of UNEB/EBDA, Campus II of the State University of Bahia, in the municipality of Alagoinhas, Bahia, Brazil (Figure 01). The material was herbarized according to standard botanical techniques (Mori et al., 1989) and deposited in the Herbarium of the State University of Bahia (HUNEB). The Atlantic Forest biome in this region is characterized by fragments and remnants, threatened by illegal logging and the advance of silviculture (Alves et al., 2023).

Figura 1 Testimonial material of *Moquiniastrum oligocephalum*, *in situ* at the EBDA/UNEB Complex, Professor's Trail, Alagoinhas, Bahia (March, 2023). Caption: A - Detail of the whitish, discolored leaves; B - Shrubby habit of *Moquiniastrum oligocephalum*, *in situ*. C and D - Capitulescence



Source: Authoral, 2023

4.3. Bioclimatic Environmental Variables

To conduct modeling procedures, the environmental variables were clipped to consider the study area, following the methodology of Austin (2007), which refers to the size of the sampling unit where occurrence data are recorded. Thus, the bioclimatic environmental variables were clipped to the specific grid corresponding to the territory of Brazil.

A set of 19 bioclimatic environmental variables were used, extracted in TIFF format from the Global Climate Data - Worldclim database (<http://www.worldclim.org/>), version 2.1 (Fick and Hijmans, 2017). Following recommendations by Chapman et al. (2005) and Phillips, Anderson, and Schapire (2006), bioclimatic variables (Table 01) composed of temperature, precipitation, isothermality, temperature seasonality, precipitation of the driest month, precipitation seasonality (coefficient of variation), and derivatives thereof were selected. These variables are relevant for studies on the spatial distribution of living organisms (Hijmans et al., 2005) and widely used in Species Distribution Modeling studies (Giannini et al., 2012; Peterson and Soberón, 2012), as well as for Ecological Niche Modeling and Environmental Suitability studies. For climatic data (i.e., temperature and precipitation), the pixel value represents the variable's value in that area.

Table 1 - Identifiers and description of bioclimatic variables

Codes	Variables
BIO ₁	Average annual temperature
BIO ₂	Average diurnal variation
BIO ₃	(monthly average - max. temp. and min. temp.)
BIO ₄	Temperature seasonality (standard deviation*100)
BIO ₅	Maximum temperature in the hottest month
BIO ₆	Minimum temperature in the coldest month
BIO ₇	Annual temperature variation (BIO ₅ -BIO ₆)
BIO ₈	Average temperature in the wettest quarter
BIO ₉	Average temperature in the driest room
BIO ₁₀	Average temperature in the hottest room
BIO ₁₁	Average temperature in the coldest quarter

BIO ₁₂	Annual rainfall
BIO ₁₃	Precipitation in the wettest month
BIO ₁₄	Precipitation in the driest month
BIO ₁₅	Rainfall seasonality (coefficient of variation)
BIO ₁₆	Precipitation in the wettest quarter
BIO ₁₇	Precipitation in the driest quarter
BIO ₁₈	Precipitation in the hottest quarter
BIO ₁₉	Precipitation in the coldest quarter

Source: *WorldClim*, 2020.

Despite the set of bioclimatic variables containing 19 variables (Table 1), a process of correlation analysis was conducted prior to modeling procedures to reduce collinearity effects in the final model. For the selection of variables to be used, the Variance Inflation Factor (VIF) technique was applied with a cutoff factor of 10; variables with VIF values less than or equal to this threshold were considered to have lower correlation within the set, while those with VIF values greater than 10 were discarded. Thus, through VIF calculation, variables with lower collinearity in the set were selected (Warren et al., 2014; Naimi and Araújo, 2016; Lima and Marchioro, 2021).

Based on the analysis of VIF values, variables with the lowest collinearity level were selected for the modeling process (Naimi and Araújo, 2016): BIO3 (Isothermality); BIO4 (Temperature seasonality); BIO7 (Annual temperature range); BIO8 (Mean temperature of wettest quarter); BIO9 (Mean temperature of driest quarter); BIO15 (Precipitation seasonality); BIO17 (Precipitation of driest quarter); BIO18 (Precipitation of warmest quarter); and BIO19 (Precipitation of coldest quarter), with the manual inclusion of BIO1 (Annual mean temperature) and BIO12 (Annual precipitation) due to their biological relevance.

The analysis and selection of the final variables for environmental suitability modeling were conducted using the *usdm* package (Naimi et al., 2014), also implemented in R 4.3.0 (R Core Team, 2023), with a spatial resolution of 2.5 arc-minutes (approximately 5 km² at the equator).

4.4. Ecological Niche Modeling

To structure the routine of Ecological Niche Modeling and quantify Environmental Suitability, the maximum entropy algorithm (MaxEnt) was utilized, an efficient method for making inferences based solely on spatial presence data of the species in a given area (Phillips et al., 2006).

MaxEnt was chosen for training the ecological niche model of *Moquiniastrum oligocephalum* in Brazil due to its ability to construct models without the need for absence points (Phillips et al., 2017), being frequently applied in studies of ecology, biogeography, conservation, among others. MaxEnt finds the distribution of maximum entropy probability (i.e., closest to uniform) subject to a set of constraints, which should correspond to the averages of the environmental variables observed in the occurrence data set that composes the sample (Phillips et al., 2006; Pearson et al., 2007; Elith et al., 2011).

The MaxEnt algorithm was used in its standard configuration as per the literature: random selection of 10,000 background points across the study area, with cross-validation replication method structured in 30 replicas. For training and testing the models, occurrence data were divided into two sets containing 70% (training) and 30% (testing) of the occurrence data, respectively.

Model validation is an area of research where new techniques have been developed and information has been disseminated through recent publications, such as statistical validation commonly recommended in the literature using AUC (Area Under the Curve) and TSS (True Skill Statistics) metrics (Thuiller et al., 2009), considering AUC values ≥ 0.7 and TSS ≥ 0.4 (Allouche, 2006; Buisson et al., 2010). Based on this, final ensembles can be produced by selecting models based on TSS, maximizing specificity and sensitivity of the models.

Due to internal differences between each generated model, a single map of consensus environmental suitability (ensemble) was created for the studied species, built from selecting the best models using a threshold (TSS ≥ 0.4), maximizing specificity and sensitivity of the adjusted models for the species (Marmion et al., 2009).

To analyze the influence of extreme points in the dataset, two maps of environmental suitability for Brazil were produced, one considering the inclusion of an extreme point (state of Roraima) and another disregarding this occurrence point. The model was adjusted for the Brazilian territory considering the species distribution, and subsequently projected for the state of Bahia (study area with the highest number of occurrence records). For ensemble creation,

consensus maps were combined into a single final map using the raster package (Hijmans, 2022), with weighted averaging to structure the final projection. The scale of environmental suitability represents the percentage of environment suitability for the species, considering the gradient represented by the model, measured from 0 to 100%.

Following modeling, to assess and statistically validate the model, the Receiver Operating Characteristics (ROC) curve technique was applied by plotting sensitivity on the y-axis and specificity on the x-axis for the cutoff limits. The ROC methodology is commonly used to evaluate the quality of models generated by MaxEnt (Phillips et al., 2006).

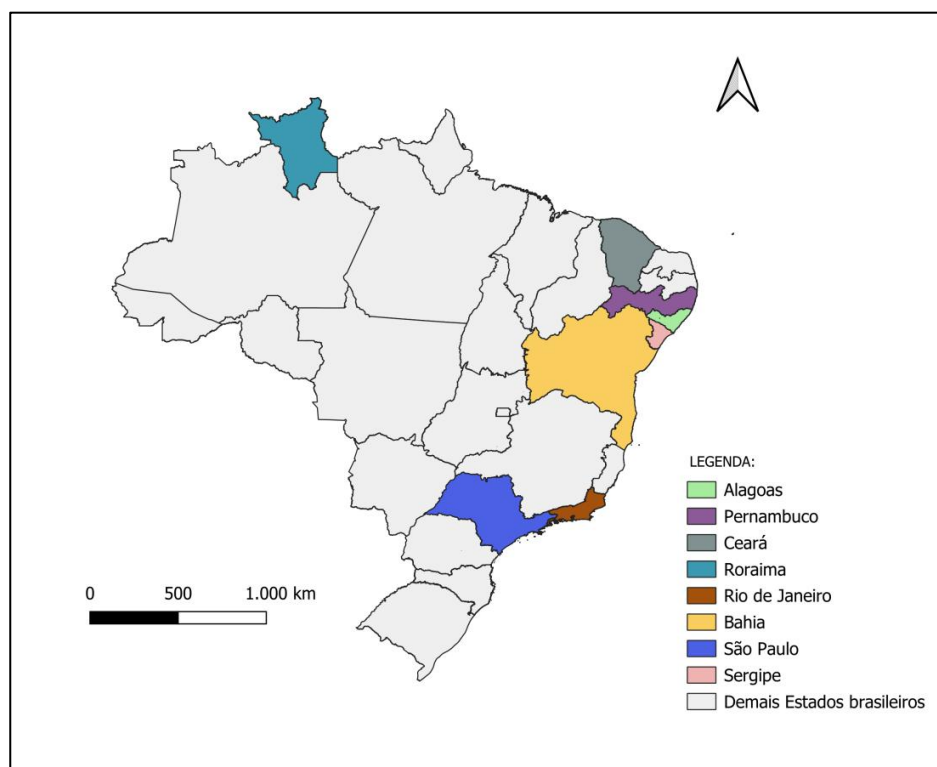
For model accuracy assessment, the Area Under the Receiver Operating Characteristic Curve (AUC) was calculated. This analysis evaluates the model's performance across all possible cutoff limits, generating a single value representing the area under the curve (AUC). It is typically determined by connecting the points with straight lines, and the area value is calculated using the trapezoid method (Phillips et al., 2006). It can be used for comparisons between different algorithms (Phillips, Anderson, and Schapire, 2006). Thus, regardless of the prior choice of a specific cutoff limit, the curve is constructed from various cutoff limits (each cutoff limit corresponds to a point on the curve).

The AUC value ranges from 0 to 1, where an AUC value of 1.0 indicates perfect discrimination. Values closer to 1 indicate higher model performance, while values around 0.5 indicate low performance (Elith et al., 2006). The closer the AUC value is to 1.0, the further the model's result is from random prediction (Phillips, Anderson, and Schapire, 2006), meaning an area equal to 1 represents the best model performance (Phillips et al., 2006; Phillips and Dudik, 2008). AUC is interpreted as the probability that the model will correctly classify a randomly chosen presence location versus any location in the landscape (Merow et al., 2013). Statistical validation was performed by calculating AUC and TSS values as described by Allouche (2006) and Buisson et al. (2010). The selection of the best models considered a TSS value ≥ 0.8 , maximizing model sensitivity and specificity.

5. RESULTS AND DISCUSSION

5.1. Geographical distribution of *Moquiniastrum oligocephalum* (Gardner) G. Sancho in Brazil For Brazil, the presence of the species under study in the states of Bahia, Ceará, Pernambuco, Sergipe, Alagoas, Rio de Janeiro, São Paulo and Roraima was verified in the databases (Figure 2). The range of occurrence of the species is necessary in relation to the occurrence of *M. oligocephalum* in Brazil, since there was a record only for the states of Bahia, Ceará and Pernambuco, according to the Flora e Funga do Brasil website, with the addition of the occurrence of the species for the states of Sergipe and Alagoas in the Northeast region and the states of Rio de Janeiro, São Paulo for the Southeast region and for the North region, Roraima.

Figure 2: Brazilian states with records of *Moquiniastrum oligocephalum* (Gardner) G. Sancho (*speciesLink* and GBIF databases, 2023)



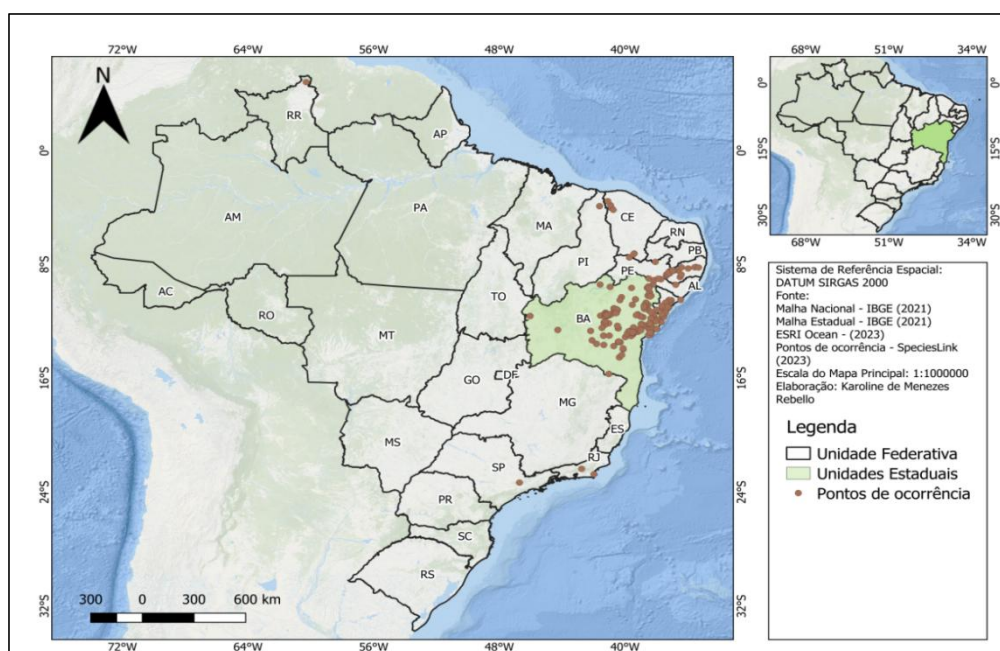
Source: IBGE Mesh, 2022. Survey data, 2024.

In north-eastern Brazil, the species predominantly occurs in drier environments with less frequent irregular rainfall. The species is usually found in arboreal habit, predominantly in environments in the phytogeographic domain of the Caatinga and Atlantic Rainforest in the northeast, as seen in the distribution map for the state of Bahia (Figure 4), where the vegetation forms a mosaic according to the type of soil and humidity.

According to Flora and Funga do Brasil (2023), *M. oligocephalum* is distributed exclusively within the phytogeographic domains of the Caatinga and Cerrado in Brazil. However, the species has been recorded in the GBIF and *speciesLink* databases occurring also in the Atlantic Forest, predominantly in the Northeast region of Brazil (Figure 2). Considering observed occurrence restrictions, it is identified that updates to the geographical distribution and phytogeographic domain in these databases are necessary to support conservation strategies for the species in Brazil.

From the information available on *speciesLink*, out of the 343 georeferenced occurrence records found, which were filtered for Brazil (Figure 3), 241 correspond to the state of Bahia, as presented in Figure 4. In Bahia, the species is more predominantly distributed compared to other states, especially in areas within the Caatinga phytogeographic domain and Caatinga/Atlantic Forest transition areas, influenced by seasonal bioclimatic factors of precipitation and temperature characteristic of the Caatinga region. The species was recorded in approximately 56 municipalities in Bahia (Figure 4) according to the *speciesLink* and GBIF platforms. In other Brazilian states, the records were: 39 for Sergipe; 38 for Pernambuco; 10 for Ceará; 8 for Alagoas; 3 for Rio de Janeiro; 1 for São Paulo; and 1 for Roraima. In the Brazilian territory, *speciesLink* identified 109 records for Atlantic Forest areas, 209 for Caatinga, and 5 for Cerrado. The Northeast region showed a higher number of occurrence records, mainly in Bahia, Pernambuco, and Sergipe.

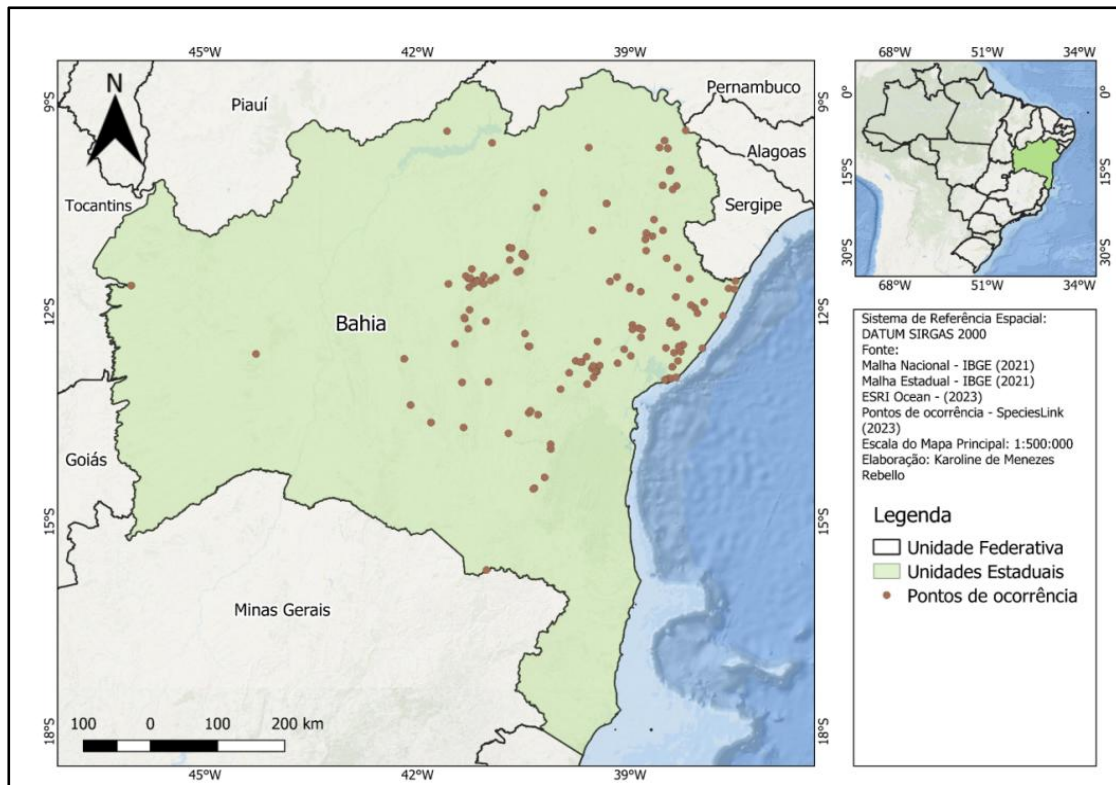
Figure 3 - Distribution map of *Moquiniastrum oligocephalum* Gardner G. Sancho for Brazil.



Source: Research data, 2024 (*SpeciesLink* and GBIF databases).

It occurs in northeastern Atlantic Rainforest environments, where there is a greater influence from the Caatinga, such as the Atlantic Rainforest on the state's northern coast (northern portion of the east coast). There is less suitability (Figure 5) for environments corresponding to the Brazilian Cerrado, where it occurs less frequently according to the databases analyzed.

Figure 4 - Distribution map of *Moquiniastrium oligocephalum* Gardner G. Sancho in the state of Bahia, Brazil.



Source: Research data, 2024 (*SpeciesLink* and GBIF databases).

According to some authors, the Asteraceae family is more prevalent in Cerrado formations and occurs less frequently in areas of Atlantic Forest, Caatinga, restinga, marshes and high altitude forests (Hind and Miranda 2008; BFG, 2015; Roque, 2016). However, this study updates this information about the species' domain of occurrence to areas predominantly in the Caatinga and Atlantic Forest phytodomains in the Northeast. The caatinga is the main vegetation formation in Bahia (SEI, 2009) and is distributed over almost the entire northeastern and central extension of the state (Queiroz et al. 2005).

M. oligocephalum was recorded in the database as occurring in the Espinhaço Chain, a region formed by a series of mountain ranges stretching for around a thousand kilometers in a north-south direction, with its northern limit at Serra da Jacobina, which cuts through the municipality of Jacobina in the state of Bahia. To the south, it reaches the Serra de Ouro Branco, in the state of Minas Gerais (Giulietti and Pitani, 1988). According to Roque et al. (2016), in the municipality of Mucugê in Bahia, *M. oligocephalum* has been collected in areas of Cerrado and campos gerais, such as the locality of Fazenda Caraíbas. Amorin and Bautista (2016) point to *M. oligocephalum* occurring in bloom, in the municipality of Jeremoabo, in the Raso da Catarina, an ecoregion characterized by a marked seasonal climate, hot and dry, of dense shrub-type Caatinga. It comprises portions of the states of Pernambuco and Bahia, and is delimited by the

Tucano-Jatobá sedimentary basin, in the central-eastern portion of the Caatinga domain (Velloso, Sampaio and Pareyn, 2002).

According to the *speciesLink* and *GBIF* databases, there are 40 occurrence records of the species in the state of Sergipe, distributed across more than 13 municipalities. These records show the species typically occurring with arboreal habit in environments that combine vegetation composition characteristic of Caatinga and forests influenced by both Caatinga and Mata Atlântica. The contact between the Mata Atlântica and the semi-arid Caatinga through the agreste establishes a zone of peculiar phytosociological transition, where species from both phytodomains coexist (Carvalho; Villar, 2005; Vicente et al., 2005, Silva et al., 2019).

Occurrences of *M. oligocephalum* are recorded in the Conservation Unit, Parque Nacional Serra de Itabaiana (PARNA), located in the central region of Sergipe (10°40' S, 37°25' W), about 40 km from the coast (Silva, Prata, and Mello, 2019). The PARNA is considered extremely important for the conservation of Mata Atlântica flora (Mendes et al., 2010; *speciesLink*, 2023). It encompasses the municipalities of Areia Branca, Itabaiana, Laranjeiras, Itaporanga D'ajuda, Malhador, and Campo do Brito (Dantas and Ribeiro, 2013).

In this ecotonal region, a considerable number of species common to the Caatinga areas stand out, although they are not endemic to this biome (Mendes, Gomes, and Alves, 2010). There are transitional zones of phytogeographical domains containing endemic species of Mata Atlântica and Caatinga (Vicente et al., 2005, Silva, Prata, and Mello, 2019). It is considered an ecological refuge due to specific substrate formation conditions (Dantas et al., 2010) and harbors a considerable number of species common to the Caatinga areas, despite not being endemic to this biome (Mendes et al., 2010). Local biodiversity is impacted by anthropogenic activities such as recreational trail use (Oliveira, 2008), logging, sand extraction, hunting, wildfires, and waste dumping (Sobral et al., 2007).

According to the *speciesLink* and *GBIF* databases, there are 37 occurrence records of the species in the state of Pernambuco, occurring predominantly in typical steppe vegetation of Caatinga in the state. From a total of 35 records consulted in the *speciesLink* database, these occur in shrub-arboreal Caatinga environments with clay-sandy soil. The remaining 2 records are from Pernambuco's Atlantic Forest. For example, *M. oligocephalum* occurs in the Environmental Protection Area Chapada do Araripe to the northwest of the state and more frequently in the central region, with records in the Catimbau National Park near the Kapinawá indigenous land, in the municipalities of Buípe and Tupanatinga. The Chapada do Araripe has areas where Caatinga and Cerrado zones mix, and there is distinct vegetation known as "carrasco" (Araújo et al., 1998; Giulietti et al., 2002). According to research by Athiê-Souza et

al. (2019), although some species have been listed as occurring in other states, out of the 613 species listed for the Catimbau National Park, 34 are new records for the state of Pernambuco, including *Moquiniastrum oligocephalum*.

In Pernambuco, there are important state and federal conservation areas established to ensure the preservation of Brazil's biological diversity (Leão et al., 2011). Among the federal ones are the Catimbau Valley National Park and Chapada do Araripe (CPRH, 2014). There are two records of *M. oligocephalum* occurring at altitudes between 889 to 991 meters in Catimbau, Buíque municipality, an important priority conservation area in the Caatinga phytodomain. Geologically, this park is part of the Jatobá sedimentary basin and is characterized topographically by low mountainous elevations (800 to 1000 m altitude) and open valleys with steep slopes (Rodal et al., 1998, SIGEP 2010). Regarding vegetation, typical of Caatinga, it also presents influences from Brazilian phytodomains of Atlantic Forest and Cerrado (Sales et al., 1998, IBAMA, 2009).

The state of Ceará covers an area of approximately 148,894.757 km², composed of 184 municipalities (IBGE, 2019). Distributed in Ceará, 10 occurrence records of the species were found, located in 5 municipalities: Crato, Viçosa do Ceará, Ubajara, São Benedito, and Guaraciaba do Norte. Remarkably, near the border with the state of Piauí in the northwest portion of the state, 8 of these occurrences of the species, including the region of the Federal Conservation Unit known as Ubajara National Park and Environmental Protection Area Serra da Ibiapaba, both located in the Ibiapaba Plateau. In addition, two records were observed in the south of the state, near the border with Pernambuco. These are located in the Federal Conservation Unit known as Chapada do Araripe Environmental Protection Area, in a region near the municipalities of Crato and Juazeiro do Norte. The Ibiapaba Plateau is the highest-altitude brejo in the westernmost location, bordering the state of Piauí. The region presents microclimates influenced by various environmental factors and species richness, which is benefited by the proximity to the coast and orographic rains (Bezerra et al., 1997; Ivanov et al., 2022).

The vegetation composition of Ceará has different phytogeological units (Figueiredo, 1997; IBGE 2012; Loiola et al., 2020), based on climate, soil types, as well as lithological and topographic characteristics. According to Lima et al. (2000), large geo-environmental compartments are recognized: Coastal Region; Fluvial Plains; Semi-arid Region; Mountainous Region. In southwestern Piauí and on the border of Ceará, there is a complex ecotone where typical Caatinga vegetation meets Cerrado, and various communities occur in these transition zones (Andrade-Lima 1978, Emperaire 1983, Oliveira et al. 1988, Araújo et al., 1999).

The state of Alagoas covers an approximate area of 27,767 km² (IBGE Resolution No. 5 of October 10, 2002). It borders to the north and west with the state of Pernambuco, to the south with the states of Sergipe and Bahia, and to the east with the Atlantic Ocean. According to IBGE (2010), the mesoregions of the state of Alagoas comprise the East, Agreste, and Sertão Alagoano. The east of Alagoas is the largest region in territorial area and includes the Coast and Zona da Mata (North and South). The Agreste is a transition region between wet and dry zones. The Sertão corresponds to areas with arid and semi-arid climatic characteristics.

According to consultations in *speciesLink* and *GBIF*, there are eight occurrence records of the species for the state of Alagoas. These occurrence points are distributed in regions in the center of the state and near the border with the state of Pernambuco. Four records are located in Caatinga vegetation composition, as found in the Serra da Boa Vista region, where granite rocky outcrops with rocky and shrub-arboreal dense Caatinga vegetation were observed in the municipality of Mata Grande.

Other records in the state of Alagoas are located in the municipality of São José da Laje (Serra das Guaribas) and also in the Biological Reserve Pedra Talhada (municipality of Quebrangulo). *M. oligocephalum* occurs in shrub-arboreal habit in fragments of Atlantic Forest in Serra das Guaribas, where part of the Pedra Talhada Biological Reserve is located.

According to Studer (2015), the Pedra Talhada Biological Reserve is considered one of the most important fragments of Atlantic Forest in the northeastern agreste. The forest is characterized as a high-altitude brejo, submontane ombrophilous forest formation (Studer, 1985; Thomas & Barbosa, 2008), where topographical characteristics directly influence rainfall, which is maintained due to the local climate, which is wetter than that of neighboring regions (Studer et al., 2015). It is located in the interior of Alagoas and Pernambuco and preserves important stretches such as the Pernambuco Interior Forests, one of the ecoregions defined by WWF in the Northeast region.

In this protected high-altitude brejo area, there is a submontane ombrophilous forest formation (Studer, 1985; Thomas & Barbosa, 2008), and topographic characteristics directly influence local rainfall, which is maintained due to the climate being wetter than neighboring regions. It is a transition region with Caatinga.

The rainfall regime in the state is directly related to configurations of large-scale atmospheric and oceanic circulation over the tropics. Annual rainfall averages vary from 2,000 mm on the coast to 400 mm in the *sertão*. Thus, distinct areas of the state exhibit different environmental and climatic aspects, such as irregularities in rainfall and little seasonal variation in solar radiation, photoperiod, and air temperature (Barros et al., 2012).

In the southeastern region of Brazil, there are records of *M. oligocephalum* occurring only in the states of São Paulo and Rio de Janeiro. Two occurrence records of the species were found in the mountainous region of the state of Rio de Janeiro. One record was located at 1922 meters altitude in an area known as Morro Branca de Neve, located in the municipality of Teresópolis and included in the Conservation Unit Três Picos State Park. The other record was in an environment characterized as semi-deciduous forest on tabuleiros, in Praia da Gorda, municipality of Armação dos Búzios.

The Serra do Mar is characterized as a geomorphological macro-unit of great importance on the south-southeast Atlantic coast of Brazil, comprising a diversified set of escarpments and plateaus (Freitas et al., 2017). The mountainous region of the state of Rio de Janeiro, where Três Picos State Park is located, consists of granitic and gneissic rocks belonging to the Eastern Terrain of the Ribeira Belt (Heilbron et al., 2000; Heilbron and Machado, 2003). According to Freitas et al. (2017), Três Picos State Park presents conditions of difficult accessibility due to geomorphological factors, such as high altitudes and steep escarpments. Due to the peculiarities of this region, which presents characteristics of mountainous relief, constant planning and monitoring are necessary for the management process (occupation and land use), as it is a priority area for conservation in the state.

Armação de Búzios is located in the region of coastal lowlands, covering an area of 71.7 km² (IBGE, 2010) and borders the municipality of Cabo Frio and the Atlantic Ocean. It presents landscapes of great geological and biological diversity, with ecosystems such as forests, restingas, rocky shores, coral reefs, mangroves, and sandy beaches. Additionally, it holds environmental, social, and cultural importance, as it also houses traditional populations (Barbosa, 2003; Dantas et al., 2009).

For the state of São Paulo, there is a single record in Juquery State Park, in the municipality of Franco da Rocha. Juquery State Park, designated as a priority conservation area, was created by State Decree No. 36,859 of 06/05/1993, originating from Juquery Farm lands (São Paulo, 1993). According to Monteiro (1973), in the semi-mountainous landscape of the area where Juquery State Park is located, the climatic feature is controlled by equatorial and tropical masses, generating tropical climates with alternating dry and wet seasons, according to Köppen's classification (1948). The altimetric gradient ranges from 730 to 950 meters, and the relief is characterized as a "sea of hills" belonging to the Serrania de São Roque, in the Atlantic Plateau region (Ab'Saber, 1978; Dantas, 1990).

The full conservation unit of Juquery State Park - PEJ stands out in the domain of the sea of hills of the Atlantic Plateau, according to Brazil (2006), where the municipalities of Franco da

Rocha and Caieiras are located in the São Paulo metropolitan region. In this region, there are remnants of preserved Cerrado biome, with relic patches, which possibly represents a broader distribution of Cerrado in the past (Ab'Saber, 1963, 2003).

The platforms consulted indicate a single record for the state of Roraima, which is located in an ecoregion known as Guiana Savanna, near the municipalities of Pacaraima and Uiramutã, on the border with Venezuela. This ecoregion is one of the units considered by Borsato et al. (2015) in their study of terrestrial ecoregions of Brazil, based on data from the Ministry of the Environment (MMA) and Wildfinder (WWF).

The states with the highest occurrence records were in the Northeast region: Bahia, Sergipe, and Pernambuco, especially in the state of Bahia. This predominance of the species indicates that the Northeast region is the center of dispersion of the species in Brazil. No records were found in the databases in the Central-West and Southern regions of Brazil, and the few records of the species for the Southeast region, in the mountainous region of the Rio de Janeiro coast and in São Paulo, show the need to expand sampling efforts and studies in the states of the Southeast region and Roraima to confirm the occurrence and distribution of the species in these states. From this study, therefore, it is known that it is necessary to carry out better analyses and invest in floristic surveys, as well as field expeditions where *Moquiniastrum oligocephalum* can be collected in various Brazilian states.

Although it is somewhat suited to occurring in an Atlantic Forest environment, it is a species that probably does not need high rainfall to survive and there is a preference for Caatinga areas, since it has adaptive morphophysiological characteristics that allow it to survive in these environments. Thus, even though it occurs in forest environments, it is possible to see that it is more suited to areas influenced by the Caatinga, i.e. those with low rainfall throughout the year.

The species has been recorded in protected areas such as Catimbau National Park, Morro do Chapéu State Park, Pedra Talhada, Junquerey Park, Serra do Mar and others. Regarding the ecoregions of the Caatinga Biome, there are records of the species occurring in the Chapada Diamantina, Raso da Catarina and the Ibiapaba-Araripe Complex. Even though it has been recorded in some protected environmental areas, there is still a lack of floristic surveys in these and other areas that have not yet been studied.

This study draws attention to the need for investment in projects aimed at floristic studies with the species, for the states of the northeast region with few or no records, especially Paraíba and Rio Grande do Norte.

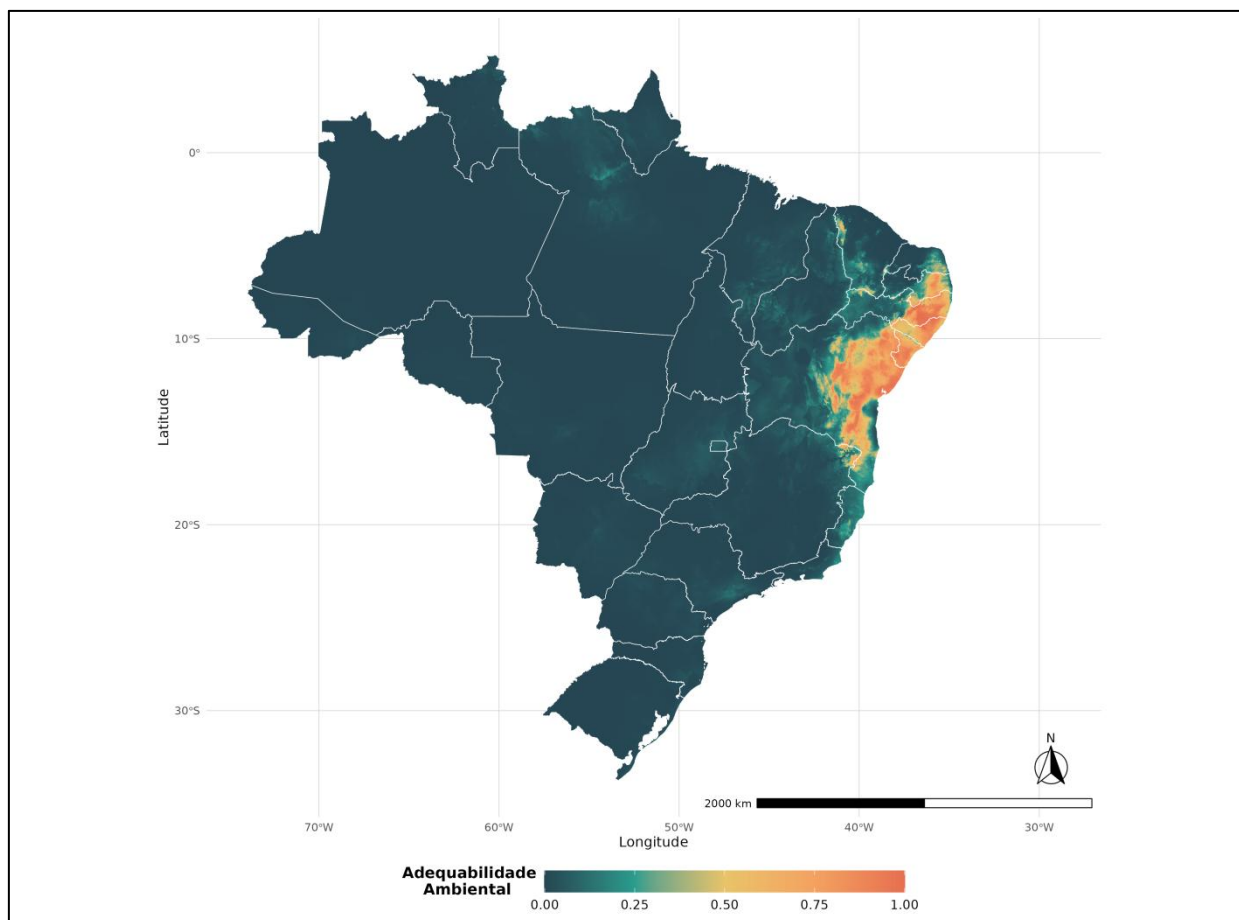
5.2. Ecological Niche Model of *Moquiniastrum oligocephalum* (Gardner) G. Sancho for Brazil

The modeling conducted corroborates information about the species distribution and indicates the Brazilian Northeast, especially the state of Bahia (Figure 6), as the center of dispersion for the species in Brazil. The probability of species occurrence can be estimated based on environmental suitability, with the percentage scale indicating how suitable or unsuitable the environment is for the species occurrence (representing potential geographic niches it may occupy).

Thus, the modeling conducted (excluding the extreme point in Roraima from the analyses) allowed visualizing regions with higher environmental suitability for Brazil. For statistical validation of the models, values of $AUC \geq 0.7$ and $TSS \geq 0.8$ were considered; the results obtained in the modeling showed $AUC = 0.97$ and $TSS = 0.87$, which are considered valid for projection.

The areas indicated with higher percentages (suitability $\geq 50\%$) of environmental suitability (Figure 5) for the occurrence of *Moquiniastrum oligocephalum* correspond to the northern portion of the East Coast, Brazilian Northeast, including mainly areas of the geographical domains of Caatinga and Mata Atlântica, predominantly in the state of Bahia. However, for the Central-West, South, North, and Southeast regions, the percentage of environmental suitability was lower (around 30% to 00%), indicating the need for updating databases and further investigation as an incentive for more field expeditions, where the species can be collected and taxonomically identified accurately.

Figure 5 - *Ensemble* modeling of *Moquiniastrum oligocephalum* (Gardner) G. Sancho projections for the Brazilian territory based on the occurrence matrix

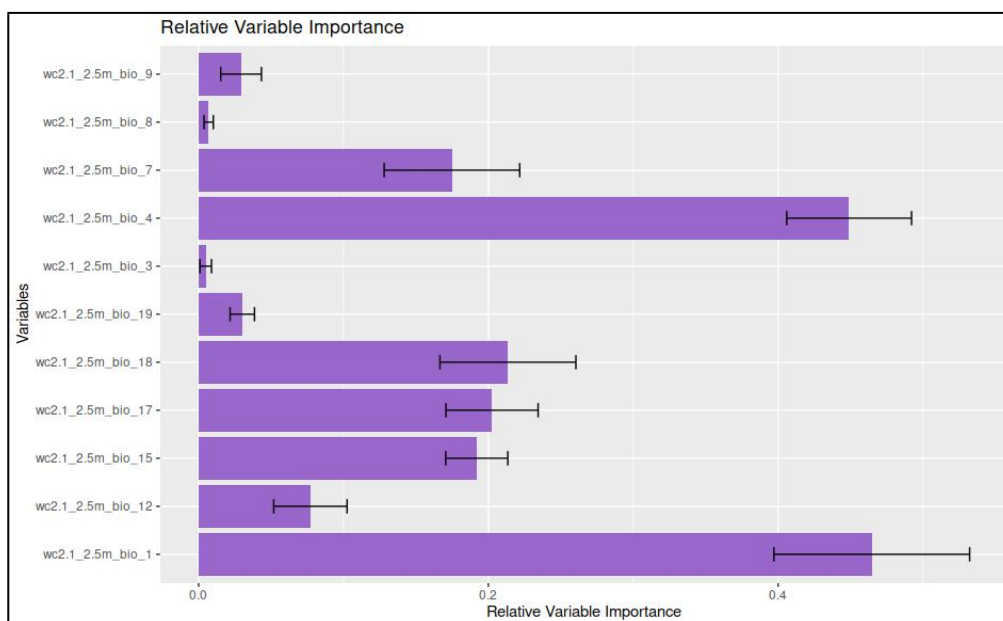


Source: Survey data, 2024

The model is satisfactory for analyzing the distribution of *M. oligocephalum* habitat suitability. Thus, it was possible to better understand that the species develops preferentially in environments with marked environmental seasonality, where there is periodic regular rainfall and temperatures characteristic of a tropical to semi-arid climate (as seen in strips comprising the Caatinga and the Atlantic Rainforest).

The relative importance values of the variables with the greatest contribution to Ecological Niche Modeling were: BIO1 - average annual temperature (46.5%); BIO4 - temperature seasonality (44.9%); BIO18 - precipitation in the hottest quarter (21.4%); BIO17 - precipitation in the driest quarter (20.3%); BIO15 - precipitation in the wettest quarter (19.2%) (Figure 6). This data shows the direct influence of these variables on modeling (Figure 5) and the limits of the geographical distribution of the species.

Figure 6 - Relative importance of bioclimatic environmental variables in the Environmental Suitability Modeling of *Moquiniastrium oligocephalum* (Gardner) G. Sancho



Source: Research data, 2024.

The modeling (Figure 5) represents, in an approximate manner, environmental conditions suitable for the occurrence of the species and presents relevant biological sense, corroborated by information regarding the investigated geographic distribution in databases. It was verified the important influence of highlighted bioclimatic variables: BIO1 - mean annual temperature (46.5%); BIO4 - temperature seasonality (44.9%); BIO18 - precipitation in the warmest quarter (21.4%) for the development and occurrence of the species in the areas shown on the species distribution map for Brazil (Figure 3). Regarding the influence of precipitation (represented by variable BIO18) and significant annual temperature fluctuations (BIO07), spatial-temporal variabilities can be perceived, where the mechanism of precipitation formation is directly related to the accumulation of water vapor volume, condensed in the air, as stated by Tucci (2000).

Amissah et al. (2014) identify precipitation as a key factor in the distribution of tree species. Additionally, according to Albuquerque et al. (2012), precipitation (rainfall) is considered the most controlling and dominant climate element for life in dry lands, such as in the Caatinga biome, where rain initiates and concludes several eco-physiological processes such as phenology and seed germination. According to Carvalho e Silva (2006), atmospheric humidity, especially due to evapotranspiration, and the mechanism of air cooling (due to the ascent of humid air) are some of the elements necessary for precipitation formation.

According to Frankie et al. (1974), Gentry (1982, 1988), and Bullock (1995), one of the main ecological differences between dry and wet tropical forests is related to the difference in the amount and seasonality of annual precipitation. Therefore, the relevance of environmental factors such as temperature seasonality and precipitation throughout the year (especially after

periods when temperatures are higher), which allows for the increase in necessary humidity for species flowering, followed by a drier period, allowing for fruit development and dispersion.

The species has characteristics that favor dispersion in different environments where it can be found. For example, there is the biological characteristic of the species to produce dry fruits of the cypsela type, with a tufted dispersal structure (pappus), which are dispersed by anemochoric events (influence of the movement of air masses) after periods of significant rainfall and suitable warm temperatures during the reproductive period.

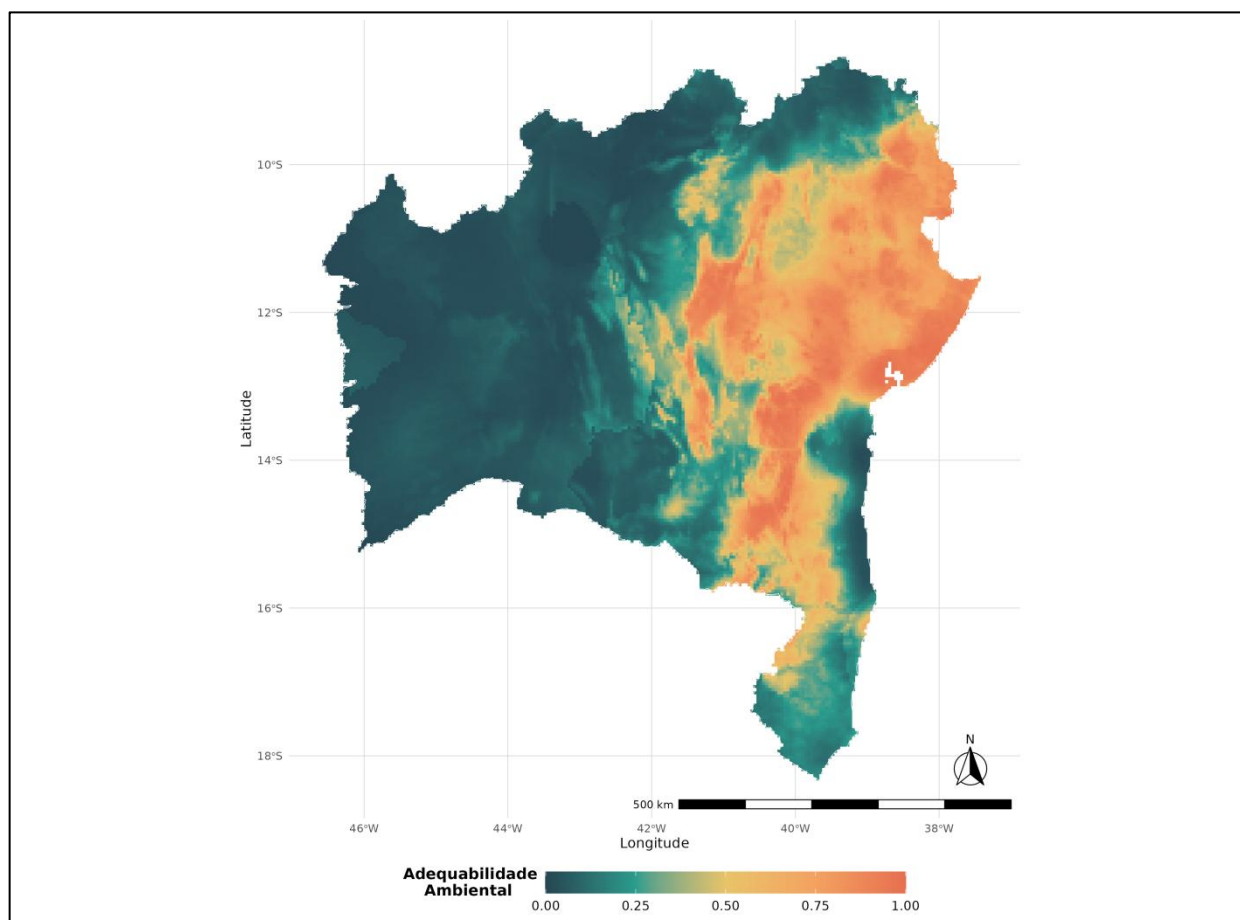
Studies in the ecology of plant species dispersion constitute an important foundation for understanding the structure and functioning of forest communities in the Neotropics (Gentry, 1983). Dispersion allows for the establishment of new individuals in a population and/or colonization of new sites, influencing the floristic composition and functioning of forest communities. In dry environments, anemochory is the predominant syndrome (Howe and Smallwood, 1982; Lopes et al., 2010). The dispersal of anemochoric diaspores is favored during drier periods in regions with seasonal climates (Croat, 1975; Morellato and Leitão-Filho, 1990), as environmental factors such as increased temperature, along with low humidity, help in fruit opening and weight reduction, as well as in the dispersal of diaspores, with reduced vegetation cover increasing exposure to wind (Souza et al., 2012).

According to Frankie et al. (1974), Bullock (1995), anemochory is the dominant mode of dispersion, characteristic of dry tropical forests. It is generally assumed that seeds dispersed by the wind prevail in "dry forests" (Gentry, 1983, 1995). Regarding the mode of dispersion of woody plants, it is assumed that the frequency of various seed dispersal strategies differs between wetter and drier locations, both at the continental and local scales, as well as between sympatric perennial and deciduous tree species (Gentry, 1995).

The information presented as a result of the highlighted bioclimatic variables (Figure 6) corroborates the environmental preference of *M. oligocephalum*, which has adaptable phytomorphological aspects that allow it to survive especially in environments characteristic of the Caatinga and Atlantic Forest under semi-arid influence, where there is marked seasonality and low rainfall. The species was recorded occurring less frequently in Cerrado environments.

According to the projection results for the state of Bahia (Figure 6), a high environmental suitability (suitability $\geq 50\%$) for *Moquiniastrum oligocephalum* was observed in areas of the state with predominantly tropical and semi-arid climates, as well as high temperatures and high rainfall in some periods. However, it showed low values of environmental suitability (suitability $\leq 25\%$) in areas of the southern East Coast, western Bahia, and Cerrado belts (Figure 7).

Figure 7 - Ensemble projections of *M. oligocephalum* for the state of Bahia from the complete Brazilian model, based on the complete occurrence matrix



Source: Author's elaboration (2024).

The species is notably distributed across the phytogeographical domains of the Caatinga in Bahia and the Atlantic Forest, where it is concentrated in the northern portion of eastern Bahia, as well as in the transition zones between the Caatinga and the Atlantic Forest. The analysis of the generated model can support environmental conservation actions for the species in the Atlantic Forest areas of the Northern Coast and in the Caatinga of Bahia (Figure 6).

It is evident that the biodiversity in the state of Bahia is substantial, with the potential to increase further with more extensive sampling efforts, as a large part of the floristic inventories are concentrated in the southern portion of the state, in ombrophilous forests or transitional areas (Amorim et al., 2009; Thomas et al., 2009; Coelho and Amorim 2014). Additionally, according to authors such as Tabarelli et al. (2006), more than 46% of the mapped remnants of the Atlantic Forest in the Northeast are located in Bahia.

In the state of Bahia, areas with high precipitation are primarily found in the eastern part of the state (Figure 7), where remnants of the Atlantic Forest are located. The protection of significant environments poses a challenge, and the promotion of scientific actions contributes to this process (Sanquenta, 2008).

For the states of Minas Gerais, Espírito Santo, Rio de Janeiro, and São Paulo (with recorded occurrences in São Paulo and Rio de Janeiro), the modeling showed reduced environmental suitability ($\leq 40\%$) compared to the states of Bahia, Sergipe, and Pernambuco.

This research emphasizes the need to expand sampling efforts through floristic surveys, identification of areas, and management of priority conservation actions for Brazilian biomes, particularly in northeastern states such as Piauí, Paraíba, and Rio Grande do Norte, as these states, for example, contain considerable portions of Caatinga ecoregions (Appendix A), and no records of *M. oligocephalum* have been found in the consulted databases.

In the northern region of Brazil, there is a predominance of Amazonian forests. In the state of Roraima, reduced environmental suitability was observed (suitability $\leq 25\%$), indicating that this is not the preferred niche for the species to occur. Regarding the single occurrence record found in the databases, which was disregarded in the latest modeling analyses, it is possible that the species was introduced in the region, as it is known that the species is not endemic to northern Brazil.

These findings corroborate what is currently known in the literature regarding the global distribution of the genus *Moquiniastrum*. Following the publication by Sancho et al. (2013), there was a nomenclatural rearrangement, elevating *Moquiniastrum* to the rank of genus. The species are limitedly distributed to South America (Appendix B) and occur mainly in Brazil (Sancho, 2000; Funk et al., 2014). However, despite efforts, there is a scarcity of studies on the taxonomy, ecology, and geographical distribution of the group, necessitating further compilation and updating of data.

In the analyses, the preferred niche for the occurrence of the species was considered to be locations with higher percentages of environmental suitability (suitability $\geq 50\%$) as indicated in the modeling. Thus, the combination of occurrence data with selected bioclimatic environmental variables allowed for the development of a representation of the probabilistic conditions necessary for the survival of *M. oligocephalum*. Nonetheless, further analyses need to be conducted.

In the research by Carvalho et al. (2019), using machine learning algorithms, the results showed a greater influence of the environmental variable precipitation compared to temperature. A study of the distribution of ten tree species in the Caatinga region, most abundant in a sub-

basin of the São Francisco River (MG), evaluated three machine learning methods (J48 decision tree, random forest, and artificial neural networks).

According to the study by Araújo et al. (2004), this type of modeling is an important tool for selecting priority areas for conservation, reinforcing the importance of maintaining the species for the preservation of the environments where it occurs. Thus, the data presented on the geographical distribution of *M. oligocephalum* enable the expansion of biogeographic studies necessary for the development of conservation strategies for species in the local flora.

5.3. Priority Areas for Species Conservation in Brazil

Brazil is the country with the highest diversity of species in the world, spread across six terrestrial biomes, as well as the coastal and marine zones. However, the biological richness, endemism, and diverse phytophysiognomies of Brazilian biomes require more efficient tools to aid environmental conservation (MMA, 2023). Since 2004, the Ministry of Environment and Climate Change has adopted the Priority Areas for Conservation, Sustainable Use, and Distribution of the Benefits of Brazilian Biodiversity.

The Priority Areas serve as an informative tool guiding the definition of environmental public policies and providing a database with information on action priorities in each area, considering biological importance and various uses of natural resources (MMA, 2023). The tool also supports already established protected areas, such as conservation units, indigenous lands, and quilombola territories, which harbor various endangered species and special environments.

Furthermore, it supports measures to be implemented in the new Priority Areas for Biodiversity. These measures contribute to actions such as research, biodiversity inventory, recovery of degraded areas, environmental licensing, monitoring, identification of areas with potential for creating conservation units, among other actions aiming at connectivity of areas through ecological corridors, promotion of sustainable use, and environmental regularization (MMA, 2023).

According to data from various databases, *M. oligocephalum* occurs in protected areas such as: National Park (PARNA) in the Serra de Itabaiana in Sergipe; Raso da Catarina on the border between Pernambuco and Bahia; Morro do Chapéu State Park in Bahia; Pedra Talhada in Alagoas; Vale do Catimbau in Pernambuco; and the Araripe Plateau located in Ceará, Pernambuco, and Piauí. Additionally, the Second *SpeciesLink* (2024) reports the species' presence in the Serra das Guaribas in Alagoas. In Ceará, *M. oligocephalum* has been recorded in

federal conservation units like the Serra da Ibiapaba Environmental Protection Area, Araripe-Apodi National Forest, and Ubajara National Park.

Regarding biodiversity identification and conservation priority areas discussed by Olson et al. (2001), the global ecoregion map provides a useful framework for conducting biogeographic or macroecological research. These are largely based on the biogeographic realms of Pielou (1979) and Udvardy (1975). A team of biologists convened by the World Wide Fund for Nature (WWF) developed a system of eight biogeographic realms (ecozones) as part of their delineation of over 800 terrestrial ecozones worldwide.

According to Borsato (2015), an important approach for defining conservation priorities is the Terrestrial Ecoregions in Brazil. This involved overlaying the geographical delineation of the ecoregions (WWF) on vegetation cover maps showing remnants of the Atlantic Forest in Brazil (MMA, 2005). Locations with the lowest proportion of remaining vegetation are considered of utmost importance for conservation. The ecoregional biodiversity vision, as conceived by WWF-Brazil, is a planning tool aimed at guiding biodiversity conservation actions to identify priority areas to maintain a representative sample of biodiversity across an entire ecoregion (Borsato, 2015).

Highlighted ecoregions and the percentage of remaining Atlantic Forest, in descending order, are: Guianan Savanna (8.3%); Bahia Coastal Forests (11.9%); Bahia Interior Forests (14.8%); Caatinga Enclaves Moist Forests (25.6%); Pernambuco Coastal Forests (27.6%); Serra do Mar Coastal Forests (29.1%); Atlantic Coast Restingas (30.2%); Pernambuco Interior Forests (33.8%); Caatinga (57.1%); Atlantic Dry Forests (66.5%) (Borsato et al., 2015).

M. oligocephalum has been recorded in the ecoregions listed above. Verified records in the consulted databases, in the Atlantic Forest environment, indicate the need for priority conservation actions in these ecoregions and other localities in Brazil.

Significant advances in conservation have been made over recent decades, such as the enactment of Law No. 11,428 of December 22, 2006, the Atlantic Forest Law, which regulates the use of Atlantic Forest remnants. However, only 9% of the remaining forest cover is within protected areas (Ribeiro et al., 2009), and these do not necessarily correspond to areas of vegetation remnants. Protected areas do not always represent the region's biodiversity. Therefore, since it is not possible to preserve all remaining natural areas, it is necessary to prioritize them, aiming to adequately represent existing biodiversity (Durigan et al., 2009).

Despite being reduced and highly fragmented, it is estimated that the Atlantic Forest hosts about 20,000 plant species (about 35% of Brazil's species), including many endemic and endangered species. The coverage of protected areas in the Atlantic Forest has significantly

increased over the years, with contributions from federal, state, and more recently, municipal governments and the private sector. However, most native vegetation remnants remain unprotected (MMA, 2020).

The development of Brazil's main humid forest blocks, the Amazon and the Atlantic Forest, is still under discussion. These two formations were possibly connected in the past, either directly, as suggested by fossil pollen grain studies (Behling et al., 2000), or indirectly, through a tangled network of corridors and forest galleries (Oliveira-Filho and Fontes, 2000). It is believed that both originated from small independent forest formations and probably developed due to favorable climate. During the Tertiary, a xeric corridor possibly established between both forest formations, and due to climatic variations, more humid formations than those observed today would have developed.

Thus, the Atlantic Forest region was likely covered by a seasonal arboreal formation during most of the Pleistocene, with riparian forests interconnecting it with the Amazon forest during the wetter interglacial periods (Costa, 2003; Percequillo et al., 2011). Such climatic oscillations, acting in both areas and especially with numerous local and punctual variations, would have influenced the distribution of taxa and the current relationships between the two largest neotropical forest blocks (Costa, 2003).

Due to the combined effects of altitude and moist oceanic air currents entering the continent, there are enclaves of mountain forests in the Caatinga, known as Brejos de Altitude or Brejos Nordestinos. The brejos-nordestinos (or brejos-de-altitude, humid mountain ranges, enclaves), a term coined by Andrade and Lins (1964) and Andrade-Lima (1964), develop in the semi-arid region of Northeast Brazil (Câmara, 2005) and present an integration of different morphoclimatic types of the Atlantic Forest and Caatinga, mainly in the state of Ceará.

In Pernambuco, these enclaves are located much closer to the eastern northeast coast (Coimbra-Filho and Câmara 1996; MMA 1993, 2000). They constitute disjunctions of the evergreen tropical forest, with altitudes ranging from 600 to 1100 meters, occurring on the tops and upper windward slopes of mountains in the Borborema Plateau (Andrade-Lima, 1960; Rodal et al., 1998). The mountain forests in Pernambuco have been more intensively studied concerning the location and conservation of remnants (Rodal et al., 1998) and floristics (Sales et al., 1998).

M. oligocephalum has been recorded in altitude brejos only in the states of Pernambuco and Alagoas, in the Borborema Plateau. According to Velloso et al., 2002, this region encompasses Caatinga ecoregions and includes parts of the states of Rio Grande do Norte, Paraíba, Pernambuco, and Alagoas. There are rock formations with particular vegetational

compositions, contrasting with the Caatinga of the semi-arid interior regions of Northeast Brazil. Borborema has a large vegetational diversity, ranging from the low caatingas of Cariris Velhos and Curimataú in Paraíba to forests very similar to coastal ones (defined as Atlantic Forest), and the mountain forests of altitude brejos (Giulietti et al., 2002).

Many areas with rocky outcrops, such as the Pedra Talhada Biological Reserve (municipality of Quebrangulo, Alagoas), are recognized for having a highly specialized flora with high levels of diversity and endemism (Correia et al., 2021). Thus, mountainous areas deserve special attention for biodiversity conservation due to the degree of endemism found (Hind, 1995; Rapini et al., 2002), and floristic inventories are needed in these regions, as the significant relevance of the information provided is essential for recognizing the local flora of such areas and their proper conservation (Moura and Roque, 2014).

In semi-arid regions, the function of Conservation Units goes beyond biodiversity conservation, representing an essential front to curb or reduce the effects of degradation and desertification of new Caatinga areas (Barbosa et al., 2005). Seasonally dry tropical forests are densely populated and have undergone intense transformations (including desertification) over the last five centuries, increasing their vulnerability to climate change (Silva et al., 2019). A recurring challenge is to develop functional strategies that help species survive in threatened environments and understand how plant distribution will be, given climate change, as rocky outcrops or inselbergs harbor various endemic species (Gomes and Alves, 2009, 2010).

The Caatinga is considered a high priority for conservation due to the complexity of its biota and the considerable number of endemic species it harbors (Andrade-Lima, 1982; Rodal, 1992; Sampaio, 1995; MMA, 2002). This is attributed to its climate and parent rock composition (Oliveira Filho et al., 2006; Giulietti et al., 2002). Conservation efforts in the Caatinga are closely linked to combating environmental degradation, desertification, and improving the quality of life for local populations. However, it remains underprotected in conservation units (Tabarelli et al., 2000; Teixeira et al., 2021; MMA, 2022), with limited scientific literature compared to other Brazilian vegetation domains (Tabarelli and Silva, 2003).

Moreover, preservation of permanent and remaining rivers is crucial in these areas, as they harbor a significant number of rare and endemic taxa, and water availability is a limiting factor for the diverse vegetation types in the Caatinga. Conservation of these forests is particularly critical, depending on the protection of their headwaters, found in regions such as swamps or mountain forests in Borborema, Serra do Araripe, among others (Giulietti et al., 2002).

Enhanced knowledge of the geographic distribution of *M. oligocephalum* and other species across Brazil is necessary, particularly in poorly protected areas like the Indigenous Territory Kiriri (Banzaê municipality), UNEB/EBDA Vegetational Complex in the North Coast (Alagoinhas municipality), Baixa dos Quelés and Povoado Casinhas (Jeremoabo municipality), as well as less known areas within Federal and State Conservation Units and others located in states within the "Polígono das Secas" in Northeast Brazil (Andrade-Lima, 1981).

There is a pressing need for increased investment in maintaining existing Conservation Units and establishing new ones, especially in the identity territories of North Coast and Agreste of Bahia; Northeastern Semi-arid region; Metropolitan region of Salvador; Portal do Sertão; Itaparica; Sisal; Recôncavo; and Chapada Diamantina, particularly in Bahia state, to enhance preservation of threatened biodiversity and ensure adequate conservation of flora.

6. FINAL CONSIDERATIONS

Brazil is a country of continental dimensions where the increasing use of natural resources is causing a gradual loss of biological diversity at various levels. Therefore, it is of utmost importance to map and develop strategies so that the quality and quantity of data can also be evaluated.

The Northeast region can be considered as the species dispersion center. The states with the highest occurrence records were in the Northeast region: Bahia, Sergipe, and Pernambuco, especially in Bahia where significant environmental suitability was observed for the Caatinga and Atlantic Forest in the northern portion of the eastern coast of the state, and low environmental suitability in areas to the south and west of Bahia. For other Brazilian regions, the ensemble map showed reduced environmental suitability at a scale of less than 25%. No records were found in the databases for the Midwest and South regions of Brazil, and the few records of the species for the Southeast region, in the mountainous coastal region of Rio de Janeiro and São Paulo, indicate the need to expand sampling efforts.

The statistical indicators (AUC and TSS) evaluated classify the model as valid, being satisfactory in estimating the habitat suitability distribution of *M. oligocephalum*. Thus, it was possible to better understand that the species preferably develops in environments with marked seasonal variability, where there are periodic regular rains and characteristic temperatures ranging from tropical to semi-arid climates (as seen in areas encompassing the Caatinga and Atlantic Forest). This study suggests the constant updating of databases, especially in support of the work developed by Flora and Funga of Brazil, since there are records in the Caatinga, Atlantic Forest, Cerrado, and transition zones between these phyto-domains.

Although showing some suitability to occur in Atlantic Forest environments, it is a species that likely does not require high rainfall to survive. It is noted that there is a certain preference for Caatinga areas in northeastern Brazil, as it presents morphophysiological adaptive characteristics that allow it to survive in these environments. Thus, even occurring in forest environments, it is known that there is higher environmental suitability ($\geq 60\%$) for areas influenced by Caatinga, i.e., areas with reduced rainfall throughout the year.

There are still gaps regarding the knowledge and distribution of the species that need to be filled concerning the Brazilian flora. Ecological Niche Modeling and the estimation of environmental suitability corroborated the species distribution pattern, proving to be an excellent methodology for studying the biogeography of other species, thereby expanding knowledge of Asteraceae diversity and distribution in Brazil. Moreover, it provides support for the recovery of

areas considered conservation priorities, contributing to strategies in creating other priority areas for conservation and environmental preservation in Brazil.

It is known that protected areas such as conservation units, indigenous lands, and quilombola territories are fundamental for maintaining representative and viable samples of biological and cultural diversity, sheltering several threatened species in special environments, which can even serve as a starting point for defining new priority areas for conservation.

This study suggests increased investment in collection expeditions and floristic surveys of *Moquiniastrum oligocephalum* in various states of Brazil. Despite its vast extent and considering the importance of the Caatinga phyto-domain for northeastern Brazil, there is still a lack of ecological information, with a scarcity of publications. Consequently, due to being poorly studied, poorly protected, and undergoing constant processes of alteration in floristic composition and vegetation physiognomy, the landscape has been profoundly altered, and areas have been reduced to small fragments. Additionally, even with some public policies in place, they often lack coordination with the interests and knowledge of the communities they are intended to serve. Therefore, the restoration of dry forests and the adoption of best practices to prevent further degradation are urgently needed to aid in the recovery of ecosystem productivity and resilience.

This study also suggests increased sampling efforts in priority areas for the conservation of Caatinga biological diversity, which are of extreme importance and are located around some marshlands, humid mountainous areas formerly covered by forests, such as Planalto da Ibiapaba do Norte/Jaburuca, Chapada do Araripe, and in the central region of Bahia in Morro do Chapéu and Raso da Catarina. For these reasons, the perpetuity of native species from the Caatinga must be ensured through projects that use, for example, morphological descriptions, gathering information on flowering and fruiting, phytosociological and floristic data in degraded areas, the use of analytical keys, distribution maps of species, ecological niche modeling, and providing habitat information to assess conservation status.

REFERENCES

- ABARELLI, M., J. M. C. SILVA, A. VICENTE & A. M. SANTOS. Analysis of representativeness of direct and indirect use conservation units in the Caatinga: preliminary analysis. Pp. 13, 2000. in: SILVA, J. M. C. & M. TABARELLI (eds.) **Workshop Evaluation and identification of priority actions for conservation, sustainable use and benefit sharing of biodiversity in the Caatinga biome**. Petrolina, Pernambuco, Brazil.
- AB'SABER, A. N. The semi-arid morphoclimatic domain of the Brazilian caatingas. **Geomorphology** 43: 1-39. 1974.
- AB'SABER, A. Contribution to the geomorphology of the cerrado area. In: SYMPOSIUM ON CERRADO. São Paulo: **EDUSP**, p. 117-124. 1963.
- AGÊNCIA ESTADUAL DE PLANEJAMENTO E PESQUISAS DE PERNAMBUCO. Ipojuca River Basin. **Recife: Condepe/Fidem**, 64p. 2005 (Watershed series of Pernambuco, 1).
- AIELLO-LAMMENS, M. E. *et al.* spThin: an R package for spatial thinning of species occurrence records for use in ecological niche models. **Ecography**, [S. l.], v.38, n.5, p.541–545, 2015.
- ALMEIDA-JR., E. B., PIMENTEL, R. M. M. & ZICKEL, C. S. Flora and life forms in a restinga area in the northern coast of Pernambuco, Brazil. **Rev. Geog.** 24 (1): 19-34. 2007.
- ALVARES, C. A., STAPE, J. L., SENTELHAS, P. C., GONÇALVES J. L. M. & SPAROVEK, G. Köppen's climate classification map for Brazil. **Meteorol. Z.** 22 (6): 711-728. 2014.
- ALCOFORADO-FILHO, F. G. A., SAMPAIO, E. V. S. B. & RODAL, M. J. N. Floristics and phytosociology of a deciduous thorny arboreal vegetation remnant in Caruaru, Pernambuco. **Acta Bot. Bras.** 17 (2): 287-303. 2003.
- ALVES, M., A. Alves-Araújo, B. S. Amorim, A. Araújo, D. Araújo, M. T. Buriel, J. L. Costa-Lima, J. Garcia Gonzalez, G. Gomes-Costa, A. Melo, J. Novaes, S. Oliveira, E. Pessoa, T. Pontes & J. Rodrigues. Inventory of Angiosperms of the Atlantic Forest fragments of Usina São José, Igarassu, Pernambuco. *In*: Buriel, M. T., A. Melo, A. Alves-Araújo & M. Alves (eds.) **Plants of the Atlantic Forest: Guide to trees and shrubs of Usina São José (Pernambuco)**: 133-158. 2013. Quick Book.
- ALVARES, C. A. et al. Köppen's climate classification map for Brazil. **Meteorologische Zeitschrift, Stuttgart**, v. 22, n. 6, p. 711-728, 2013.
- ALMEIDA, G. S. S. **Asteraceae Dumort. in rocky fields of Itacolomi State Park, Minas Gerais, Brazil**. Doctoral Thesis. Universidade Federal de Viçosa, Viçosa. 365f. 2008.
- ALMEIDA Jr., E. B., PIMENTEL, R. M. M. & Zickel, C. S. Flora and life forms in a restinga area on the Northern Coast of Pernambuco, Brazil. **Revista de Geografia** 24: 19-34, 2007.
- ALVES, D. S.; ALMEIDA, G. S. S.; SANTANA, I. N. Lamiaceae Martinov in the municipality of Alagoinhas, Bahia, Brazil. **Paubrasilia**, Porto Seguro, v. 6, p. 14; 2023.

AMANE (Association for the Protection of the Northeast Atlantic Forest). 2014. Map of the Northeast Atlantic Forest Corridor. Available at <http://www.amane.org.br/publicacoes.asp?pag=2/>. Accessed on November 1, 2014.

AMORIM, V. O.; BAUTISTA, H. P. Asteraceae of the Raso da Catarina Ecoregion, Bahia, Brazil. Rio de Janeiro: **Rodriguésia**, v.67, n.3, p.785-794, 2016.

AMORIM, L. A., & Santos, A. M. dos. Impacts of the Last Drought Cycle on Agricultural Production in Desertification-Susceptible Municipalities in the Semi-Arid Region of Pernambuco. **Geografia** (Londrina), 33(1), 249–263. 2023.

AMORIM, A. M., JARDIM, J. G., LOPES, M. M. M., FIASCHI, P., BORGES, R. A. X., PERDIZ, R. O. & Thomas, W. W. Angiosperms in remnants of montane forest in southern Bahia, Brazil. **Biota Neotropica** 9: 313-348. 2009.

AMORIM, A. M., THOMAS, W. W., CARVALHO, A. M. V. & JARDIM, J. G. Floristics of the Una Biological Reserve, Bahia, Brazil. In: W.W. Thomas (ed.). **The Atlantic Coastal Forest of Northeastern Brazil**. Memoirs of The New York Botanical Garden, New York, pp. 67-146. 2008.

AMORIM, A. M., FIASCHI, P., JARDIM, J. G., THOMAS, W. W., CLIFTON, B. C. & Carvalho, A. M. V. The vascular plants of a forest fragment in southern Bahia, Brazil. **Sida** 21: 1726-1752. 2005.

ANDERSON, R. P.; LEW, D.; PETERSON, A. T. Evaluating predictive models of species' distributions: criteria for selecting optimal models. *Ecological Modelling*, v.162, p.211-232, 2003.

ANDERSON, R. P.; Martínez-Meyer, E. Modeling species geographic distributions for preliminary conservation assessments: an implementation with the spiny pocket mice (*Heteromys*) of Ecuador. *Biological Conservation*. Pp. 167-179. 2004.

ANDRADE-LIMA, D. Present-day forest refuges in Northeastern Brazil. in: G.T. Prance (ed.) *Biological diversification in the tropics*. Columbia University Press, New York. 245-251f. 1982.

ANDRADE-LIMA, D. Plants of the caatingas. **Academia Brasileira de Ciências**, Rio de Janeiro. 23 f. 1989.

ANDERBERG, A. A. *et al.* Compositae. In: KADEREIT, J. W.; JEFFREY, C. (eds.). *The Families and Genera of Vascular Plants*. **Springer**, Berlin, v.8, p.61-588, 2007.

ANDRADE-LIMA, D. Present-day forest refuges in Northeastern Brazil. in: G.T. Prance (ed.) **Biological diversification in the tropics**. Columbia University Press, New York. Pp 245-251 1982.

ANDRADE-LIMA, D. 1989. Plants of the caatingas. *Academia Brasileira de Ciências*, Rio de Janeiro.

BEZERRA, M. F. A.; FARIAS, G. R.; CAVALCANTI, L. H. Mixobiota of Serra de Itabaiana National Park, SE, Brazil: Trichiales. **Acta Botânica Brasilica**, Belo Horizonte, v. 24, n. 2, p. 510-517, 2010.

ANTONELLI, A. Biogeography: Drivers of bioregionalization. **Nature Ecology & Evolution**, v.1, n.0114, 2017.

ARAÚJO, H. F. P., CANASSA, N. F., MACHADO, C. C. C. et al. Anthropogenic disturbance is the main cause of vegetation changes in the Caatinga region. **Rep Sci** 13, 18440 2023.

ARAÚJO, E. L., E. V. S. B. SAMPAIO & M. J. N. RODAL. Floristic composition and phytosociology of three Caatinga areas in Pernambuco. **Revista Brasileira de Botânica** 27: 797-810. 2004.

ARAÚJO, E. L., et al. The flora of the Paque Nacional do Catimbau, Pernambuco, Brazil. **Revista Brasileira de Botânica** 23: 401-412. 2000.

ARAÚJO, E. L., E. V. S. B. SAMPAIO, M. J. N. RODAL & G. D. BARBOSA. Composition and phytosociology of a lowland thorn woodland, Cabo de Santo Agostinho, Pernambuco, Brazil. **Revista Brasileira de Botânica** 26: 237-248. 2003.

ARAÚJO, E. L., M. J. N. RODAL & J. M. C. SILVA. Analysis of the composition and structure of seasonal semi-deciduous forests in Pernambuco, Brazil. **Revista Brasileira de Botânica** 23: 187-198. 2000.

ARAÚJO, E. L., M. J. N. RODAL & E. V. S. B. SAMPAIO. Composition and structure of Caatinga vegetation in the semi-arid northeastern Brazil. **Bulletin of the Torrey Botanical Club** 125: 341-347. 1998.

ARAÚJO, A. P. (2010). Asteraceae in the Jurubatiba restinga. Rio de Janeiro: **Acta Bot. Bras.** vol.24. pp. 3-19.

ARAÚJO, H. F. P., F. A. M., & SÁNCHEZ, M. N. Vegetation and environmental factors in the Atlantic Forest - restinga vegetation interface in northeastern Brazil. 2023.

ARARIPE, H.G.A.; LOPES, J.B.; BASTOS, M.E.G. Environmental licensing aspects of shrimp farming in the APA of Delta do Parnaíba. **Ambiente & Sociedade**, v.9, n.2, p.143-173, 2006.

AUSTIN M. Species distribution models and ecological theory: a critical assessment and some possible new approaches. **Ecological Modelling**, v.200, n.1, p.1-19, 2007.

BAÑUELOS, M. J. et al. Modelling the distribution of *Ilex aquifolium* at the north-eastern edge of its geographical range. **Nordic Journal of Botany**, v.23, n.1, p.129-142, 2004.

BARBOSA, M.R.V., CASTRO, R., ARAÚJO, F.S. & RODAL, M.J.N. Strategies for biodiversity conservation and research priorities in the Caatinga biome. In: **Analysis of Biodiversity Variations in the Caatinga Biome: Supporting Regional Strategies** (F.S. Araújo, M.J.N. Rodal & M.R.V. Barbosa, eds). Ministry of the Environment, Brasília, p. 417-431, 2005.

BARBOSA, K. C. Tourism in Armação dos Búzios (RJ/Brazil): Local perceptions of city problems and prioritized guidelines for environmental management. 124 p. 2003. Dissertation (Master in Environmental Science) - Federal Fluminense University, Niterói, 2003.

BARROS, A. H. C.; Climatology of the State of Alagoas / Alexandre Hugo Cezar Barros - Electronic data. Recife: Embrapa Solos, (Research and Development Bulletin / Embrapa Solos. 32 p. ISSN 1678-0892; 211), 2012.

BELLARD, C. et al. Vulnerability of biodiversity hotspots to global change. **Global Ecology and Biogeography**, v.23, n.12, p.1376–1386, 2014.

BEZERRA, E. C., J. E. G. BEZERRA & M. F. S. MENDES. 1997. Precipitation. Pp. 22-23 in: IPLANCE (ed.) **Atlas of Ceará**. Ceará Institute of Planning Foundation, Government of Ceará, SEPLAN, Fortaleza.

BFG. Brazilian Flora 2020: Innovation and collaboration to meet Target 1 of the Global Strategy for Plant Conservation (GSPC). **Rodriguésia**, v.69, n.4, p.1513-1527. 2018.

BFG. Growing knowledge: an overview of Seed Plant diversity in Brazil. **Rodriguésia**, v.66, n.4, p.1085-1113. 2015.

BFG. Brazilian Flora 2020: Leveraging the power of a collaborative scientific network. **Taxon**. 2021.

BORGES-NOJOSA, D. M. & U. CARAMASCHI. 2000. Geographic distribution: **Aulura anomala**. **Herpetological Review** 31: 108.

BONATELLI, I. A. S. et al. Interglacial microrefugia and diversification of a cactus species complex: phylogeography and palaeodistributional reconstructions for *Pilosocereus aurisetus* and allies. **Molecular Ecology**, v.23, i. 12, p.3044-3063, 2014.

BORSATO, R.; LOYOLA, R.; LEMES, P. Ecoregions of Brazil: terrestrial and marine priorities. Curitiba: **LIFE Institute**, 2015.

BRAZIL. Federal Law No. 9,985, of July 18, 2000. Regulates art. 225, § 1, incisos I, II, III and VII of the Federal Constitution, establishes the National System of Units of Conservation of Nature and provides other provisions. Available at: http://www.planalto.gov.br/ccivil_03/leis/19985.htm. Accessed on: October 24, 2023.

BRAZIL. Decree of June 15, 2005. Creation of the Serra de Itabaiana National Park. Brasília: **Official Gazette of the Union**, Executive Branch, June 15, 2005.

BRITO, L.T. de L.; MOURA, M.S.B. de; GAMA, G.F.B. Potential of rainwater in the Brazilian Semiarid. **Petrolina: Embrapa Semiarid**, 181p., 2007.

BREMER, K. **Asteraceae, cladistics and classification**. Timber Press, Portland, 752p. 1994.

BRITO, G.R. **Application of predictive species distribution modeling as a tool for biodiversity study**. 2018. 121 p. Dissertation (Academic Master in Biosciences). – São Paulo State University (UNESP), Faculty of Sciences and Letters, Assis, 2018.

BROWN J. L. et al. PaleoClim, high spatial resolution paleoclimate surfaces for global land areas. **Scientific Data**, v.5, n.180254, 2018.

BROWN, J. H.; Lomolino, M. V. **Biogeography**. 2nd ed. Funpec, Ribeirão Preto. Pp. 692. 2006.

BROOKS, T.; BALMFORD, A. Atlantic forest extinctions. **Nature**, v. 380, n. 6570, p. 115–115, March 14, 1996.

BUISSON, L. et al. Uncertainty in ensemble forecasting of species distribution. **Global Change Biology**, v.16, n.4, p.1145–1157, 2010.

BULLOCK, S. H. Plant reproduction in neotropical dry forests. Pp 277-303, 1995. in: S. H. Bullock, H. A. Mooney & E. Medina (eds.) **Seasonally Dry Tropical Forests**. Cambridge University Press, Cambridge.

CABRAL-NETO, J.P.; FEITOSA, F.R.S.; AGUIAR, W.J.; PONTES, B.A.N.M. Socio-environmental diagnosis of the Delta do Parnaíba in the coastal region of Piauí-Brazil. In: **Environmental Knowledge: Reflections on the Society-Nature Relationship** (C.J.M. Castilho, ed.). Itacaiúnas, 2020.

CABRERA, A. L. Review of the genus *Gochnatia* (Compositae). **Revista del Museo de La Plata**, La Plata, v.12, p.1-160, Sept. 1971.

CABRERA, A. L.; WILLINK, A. Biogeography of Latin America. 2nd ed. Washington: OAS, 1980. 117 p.

CARDOSO, D.B.O.S.; FRANÇA, F; NOVAIS, J.S.; FERREIRA, M.H.S; Santos, R.M.; CARNEIRO, V.M.S & Gonçalves, J.M.. Floristic composition and phytogeographic analysis of a semideciduous forest in Bahia, Brazil. **Rodriguésia** 60: 1055-1076, 2009.

CARDOSO, D.B.O.S. & QUEIROZ, L.P. Leguminosae diversity in the caatingas of Tucano, BA: Implications for the phytogeography of the semi-arid Northeast of Brazil. **Rodriguésia** 58: 379-371, 2007.

CARDOSO, D.B.O.S. & QUEIROZ, L.P. 2011. Caatinga in the context of a metacommunity: evidence from biogeography, phylogenetic patterns and species abundance in Leguminosae. In: Carvalho, C.J.B.; Almeida, E.A.B. (eds.). **Biogeography of South America: Patterns and Processes**. Roca, São Paulo. Pp. 241-260. *Rodriguésia* 58: 379-371, 2007.

CARVALHO, C. M.; VILLAR, J. C. **Serra de Itabaiana National Park: survey of Biota**. (General and Experimental Biology - UFS) São Cristóvão: UFS,. 124 p. 2005.

CARVALHO-SOBRINHO, J.G. & QUEIROZ, L.P. Floristic composition of a fragment of Atlantic Forest in Serra da Jibóia, Santa Terezinha, Bahia, Brazil. **Sitientibus Series Biological Sciences** 5: 20-28. 2005

CAVALCANTE A.M.B.; FERNANDES, P.H.C.; SILVA, E.M. Ahead of Print. **Opuntia ficus-indica** (L.) Mill. and Climate Change: An Analysis in Light of Species

CHATTERJEE, S.; HADI, A. S. **Regression analysis by example**. 4th ed ed. Hoboken, N.J: Wiley-Interscience, 2006.

COELHO, M.M. & AMORIM, A.M. Floristic composition of the Montane Forest in the Almadina - Barro Preto axis, Southern Bahia, Brazil. **Biota Neotropica** 14: 1-41, 2014.

CONDAMINE, F. L. *et al.* Teasing Apart Mountain Uplift, Climate Change and Biotic Drivers of Species Diversification. **Mountains, Climate and Biodiversity**, Hoboken, Nova Jersey, p.257-272, 2018.

COUTINHO, L.M. Contribution to the understanding of the ecological role of fires in the flowering of cerrado species. 173 p. 1976. Thesis (Habilitation) – University of São Paulo, São Paulo.

CPRH - Companhia Pernambucana do Meio Ambiente. Tatu-Bola Wildlife Refuge. Recife: CPRH, 80 pp. 2014.

COLOMBO, A.; JOLY, C. Brazilian Atlantic Forest lato sensu: the most ancient Brazilian forest, and a biodiversity hotspot, is highly threatened by climate change. **Brazilian Journal of Biology**, v. 70, n. 3 suppl, p. 697–708, out. 2010.

COIMBRA-FILHO, A. F. & I. G. CÂMARA. The Original Limits of the Atlantic Forest Biome in Northeast Brazil. **Fundação Brasileira para a Conservação da Natureza**, Rio de Janeiro, 1996.

COIMBRA-FILHO, A. F., I. G. CÂMARA & A. B. RYLANDS. On the geographic distribution of the red-handed howler monkey, *Alouatta belzebul*, in Northeastern Brasil. **Neotropical Primates** 3: 176-179. 1995

CORREIA, J. S. *et al.* Floristic diversity of rocky outcrops in the Pedra Talhada Biological Reserve, Quebrangulo, Alagoas. *Revista Brasileira de Geografia Física*, [place of publication not specified], v. 14, n. 2, pp. 743-757, 2021.

CRAW, R.C., Biogeography and biogeographical principles. **New Zealand Entomologist**, 8:49-52. 1984.

CRISCI, J.V. The voice of historical biogeography. **Journal of Biogeography**, 28: 157–168, 2001.

CRISCI, J.V., KATINAS L. & POSADAS P. Historical biogeography: an introduction. Cambridge, Harvard University Press. 2003

CROIZAT, L. Space, time, form: the biological synthesis. Publicado pelo autor, Caracas. 1964.

DAI, X.; WU, W.; JI, L.; TIAN, S.; YANG, B.; GUAN, B.; WU, D. MaxEnt model-based prediction of potential distributions of *Parnassia wightiana* (Celastraceae) in China. **Biodiversity Data Journal**, [S. l.], v.10, p. e81073, 2022.

DANTAS, H.G.R.; LIMA, H.C.; BOHRER, C.B.A. Mapping of vegetation and landscape in the municipality of Armação dos Búzios, Rio de Janeiro, Brazil. **Rodriguésia**, v.60, pp.25–38, 2009.

DANTAS T.V.P. & RIBEIRO A.S. Characterization of the vegetation of Serra de Itabaiana National Park, Sergipe - Brazil. **Biotemas**, 23(4): 9–18, 2010.

DANTAS, M. E. Geomorphology of the State of Rio de Janeiro. Companhia de Pesquisa de Recursos Mineiros (CPRM), Rio de Janeiro, report and map, 63 p., 2000.

DANTAS T.V.P., NASCIMENTO - JÚNIOR J.E., RIBEIRO A.S. & PRATA A.P.N. Floristics and structure of shrub-tree vegetation of Areias Brancas in Serra de Itabaiana National Park/Sergipe, Brazil. **Revista Brasileira de Botânica**, 33(4): 575–588, 2010.

DANTAS, H.G.R.; LIMA, H.C.; BOHRER, C.B.A. Mapping of vegetation and landscape in the municipality of Armação dos Búzios, Rio de Janeiro, Brazil. **Rodriguésia**, v.60, pp.25–38, 2009.

DEAN, W. With broadex and firebrand: the destruction of the Brazilian Atlantic Forest. **Berkeley**: University of California Press, 1995.

DECREE No. 44,099, dated July 12, 1999. Incorporates into the Juquery State Park the specified area located in the municipality of Franco da Rocha. Official Gazette of the State of São Paulo, Executive Branch, v. 109, n. 130, July 13, 1999. Section I, pp. 2-3.

DINERSTEIN E, Olson DM, Graham DJ, Webster A L, Primm SA, Bookbinder MP, Ledec G. **An Assessment of the Conservation of Terrestrial Ecoregions in Latin America and the Caribbean**. Washington (DC): World Bank (1995).

OFFICIAL GAZETTE OF THE UNION (DOU). 2002. Decree 913/12 of December 13, 2002. No. 242. December 16, 2002.

DUTRA, G. F. et al. Protected areas in Brazil's coastal and marine management. In: MARONE, E.; RIET, D.; MELO, T. (eds). **Atlantic Brazil: a country rooted in the forest. Sea of Ideas & BioAtlântica Institute**, Rio de Janeiro, Brazil, 2010.

DURIGAN, G. et al. The vegetation of priority areas for Cerrado conservation in São Paulo State, Brazil. **Edinburgh Journal of Botany**, v. 60, n. 2, p. 217-241, 2003a.

ELITH, J. *et al.* Novel methods improve prediction of species' distributions from occurrence data. **Ecography**, v.29, p.129–151, 2006.

ELTON, C. S. **Animal Ecology**. Sidgwich and Jackson, London, 1927. Certainly! Here's the translation into formal and academic English:

EMBRAPA. **Brazil Climate Database**. 2014. Available at: <http://www.bdclima.cnpm.embrapa.br/>. Accessed on November 1, 2014.

FERNANDES, Moabe F.; QUEIROZ, Luciano P. Vegetation and flora of the Caatinga. **Ciência e cultura**, v. 70, n. 4, p. 51-56, 2018.

FERNANDES, M. F.; CARDOSO, D.; DE QUEIROZ, Luciano P. An updated plant checklist of the Brazilian Caatinga seasonally dry forests and woodlands reveals high species richness and endemism. **Journal of Arid environments**, v. 174, p. 104079, 2020.

FERRAZ, Raphael Cavalcanti et al. Phytosociological survey in a Caatinga area at the Grota do Angico natural monument, Sergipe, Brazil. **Revista Caatinga**, v. 26, n. 3, p. 89-98, 2013.

FERREIRA, G. C. Environmental modeling of tree species in the Vale do Jari, Monte Dourado, Pará using forest inventory data. 204 f. 2009 - Thesis (Doctorate) - **Instituto de Pesquisas Jardim Botânico do Rio de Janeiro/Escola Nacional de Botânica Tropical**, Rio de Janeiro.

FERRIER, S. Mapping spatial pattern in biodiversity for regional conservation planning: where to from here? **Systematic Biology**, v. 51, p. 331-363, 2002.

FICK, S.E.; HIJMANS, R.J., WorldClim 2: new 1km spatial resolution climate surfaces for global land areas. **International Journal of Climatology** v. 37, n. 12, p. 4302-4315, 2017.

FIGUEIREDO, M. A. The humid mountain ranges in Ceará and food production for the semi-arid region of Ceará. **Coleção Mossoroense** 523: 1-15. 1988.

FIGUEIREDO, M. A. 1997. Phytogeographical Units. Pp. 28-29 in: IPLANCE (ed.) Atlas of Ceará. **Fundação Instituto de Planejamento do Ceará**, Government of the State of Ceará, SEPLAN, Fortaleza.

FIGUEIREDO, M. A. & M. A. BARBOZA. Vegetation and flora in the Serra de Baturité - Ceará. **Coleção Mossoroense** 747: 1-10. 1990.

FIGUEIREDO, M. A., A. FERNANDES, M. B. DIÓGENES & S. S. OLIVEIRA. The Rubiaceae family in the Serra de Baturité - Ceará. **Coleção Mossoroense** 749: 1-11. 1990.

FIGUEIREDO, M.A. Vegetation. In: **Ceará. Atlas of Ceará**. Fortaleza: IPLAN-CE, 1997. p. 29-30.

FLORA E FUNGA OF BRAZIL. Flora of Brazil. Rio de Janeiro: JBRJ. Available at: <http://floradobrasil.jbrj.gov.br/>. Accessed June 5, 2023.

FLORA OF BRAZIL. Brazilian Flora 2023. Rio de Janeiro Botanical Garden, 2019. Available at: <http://floradobrasil.jbrj.gov.br/>. Accessed: June 19, 2023.

FOLLADOR, M. et al. Modelling Potential Biophysical Impacts of Climate Change in the Atlantic Forest: Closing the Gap to Identify Vulnerabilities in Brazil. In: **Climate Change Adaptation in Latin America**, 2018. ISBN: 978-3-319-56945-1.

FRANKIE, G. W., H. G. BAKER & P. A. OPLER. Comparative phenological studies of trees in tropical wet and dry forests in the lowlands of Costa Rica. **Journal of Ecology** 62: 881-913. 1974.

FRANKLIN, J.; MILLER, J. A. **Mapping Species Distributions: Spatial Inference and Prediction**. Cambridge University Press, 320 p. , 2010.

FREITAS, M.M.; SILVA, L.G.E; TUPINAMBÁ, M. Geomorphological domains in the Três Picos State Park-RJ and their relationship with geomorphological aspects. **Sociedade e Natureza Journal** 29 (01), p. 26; 2017

FUNK, V. A.; RICHARDSON, K. S. Systematic data in biodiversity studies: use it or lose it. **Systematic Biology**. v.51, n.2, p.303–316, 2002.

FUNK, V. A.; ZERMOGLIO, M. F.; NAZIR, N. Testing the use of specimen collection data and gis in biodiversity exploration and conservation decision making in Guyana. **Biodiversity and conservation**, v.8, p.727-751, 1999.

FOUNDATION SOS MATA ATLÂNTICA. **Atlas of Forest Remnants of the Atlantic Forest** - Period 2016-2017. São Paulo: SOS Mata Atlântica. Available at: https://www.sosma.org.br/link/Atlas_Mata_Atlantica_20162017_relatorio_tecnico_2018_final.pdf. Accessed on: July 5, 2023.

FUNK, F. A. *et al.* A Phylogeny of the Gochnatieae: understanding a critically placed tribe in the Compositae. **Taxon**, v.63, p.859–882, 2014.

GABRIEL, S.; BENNETT, C.; BA, A.M.; HENRY, S. Modeling the Suitability Index of Selected Conifers on Mambilla Plateau Taraba State, Nigeria: Implication on Planted Forest. **International Journal of Agroforestry Remote Sensing and GIS**, v.1, n.1-9, 2015.

GBIF – **GLOBAL BIODIVERSITY INFORMATION FACILITY**. Occurrences. Disponível em: <https://www.gbif.org/>. Acesso em: 05 jun. 2023.

GENTRY, A. H. Diversity and floristic composition of neotropical dry forests. *In*: Bullock, S. H., H. A. Mooney & E. Medina (eds.). **Seasonally Dry Tropical Forests** : 146-194.1995. Cambridge University Press. 1995.

GENTRY, A. H. Neotropical floristic diversity: phytogeographical connections between Central and South America: Pleistocene climatic fluctuations, or an accident of the Andean orogeny? **Annals of Missouri Botanical Garden** 69:557-593. 1982.

GENTRY, A. H. Dispersal ecology and diversity in neotropical forest communities. **Sonderband Naturwissenschaftlicher Verein Hamburg** 7:303-314. 1983.

GENTRY, A. H. Changes in plant community diversity and floristic composition on environmental gradients. **Annals of Missouri Botanical Garden** 75:1-34. 1988.

GENTRY, A. H. Four neotropical rainforests. New Haven, Yale University Press. 1990.

GENTRY, A. H. Diversity and floristic composition of neotropical dry forests. Pp 146-193. 1995. in: H. Bullock, H. A. Mooney & E. Medina (eds.) **Seasonally Dry Tropical Forests**. Cambridge University Press, Cambridge.

GOMES, D'Ávilla Ruama Fernandes Lopes. Climatic variables determine functional composition in seasonally dry tropical forests. 2019. 64f. Dissertation (Graduate Program in Ecology and Conservation - PPGEC) - State University of Paraíba, Campina Grande, 2021.

GIOVANELLI, J. G. R. *et al.* Ecological niche modeling of Phyllomedusa ayeaye (Anura: Hylidae): predicting new occurrence areas for a rare species. **Neotropical Biology and Conservation**, v.3, p.59-65, 2008.

- GIANNINI, T. C. et al. Current challenges in predictive modeling of species distribution. **Rodriguésia**, Rio de Janeiro, v.63, n.3, p.733-749, Sept. 2011.
- GIANNINI, T. C.; SIQUEIRA, M. F.; ACOSTA, A. L.; BARRETO, F. C. C.; SARAIVA, A. M.; ALVES-DOS-SANTOS, I. Current challenges in predictive modeling of species distribution. **Rodriguésia**, v.63, n.3. Rio de Janeiro, 2012.
- GIULIETTI, A. M., R. M. Harley, L. P. Queiroz, M. R. V. Barbosa, N. N. A. L. Bocage, M. A. Figueiredo. Endemic species of the Caatinga. In: E. V. S. B. Sampaio, A. M. Giuliatti & C. Gamarra Rojas (eds.). **Vegetation and flora of the Caatinga**: 11-24. p. 103-118. 2002. APNE (Northeast Plants Association), Recife.
- GOVAERTS, R. How many species of seed plants are there? *Taxon*, v.50, p.1085-1090, 2001.
- GOMES, D'Ávilla Ruama Fernandes Lopes. Climatic variables determine functional composition in seasonally dry tropical forests. 2019. 64f. Dissertation (Graduate Program in Ecology and Conservation - PPGEC) - State University of Paraíba, Campina Grande, 2021.
- GRAHAM, C. H. *et al.* New developments in museum-based informatics and applications in biodiversity analysis. **Trends in Ecology and Evolution**, v.19, p.497-503, 2004.
- GRINNELL, J. The origin and distribution of the chestnut-backed chickadee. **Auk**, v.21, p.375-377, 1904.
- GRINNELL J. The niche-relationship of the California thrasher. **Auk**, .34, p.427-433, 1917.
- GRINNELL, J. Geography and evolution. **Ecology**, v.5, p. 25-229, 1924.
- GROOMBRIDGE, B. (Ed.). Global biodiversity: Status of the earth's living resources. **London: Chapman & Hall**, 585 p. 1992.
- GROMBONE, M.T. et al. Phytosociological structure of the montane semideciduous forest of Parque Municipal da Grota Funda, Atibaia (SP). *Acta Bot. Bras.*, v. 4, n. 2, p. 47-64, 1990.
- GUISAN, A.; ZIMMERMANN, N. E. Predictive habitat distribution models in ecology. **Ecological Modelling**, v.135, p.147-186, 2000.
- GUISAN, A.; THUILLER, W. Predicting species distribution: offering more than simple habitat models. **Ecology Letters**, v.8, p.993-1009, 2005.
- GUISAN, A. *et al.* Using niche-based models to improve the sampling of rare species. **Conservation biology**, v.20, p.501-511, 2006.
- GUISAN, A.; GRAHAM, C. H.; ELITH, J.; HUETTMANN, F.; The NCEAS Species Distribution Modelling Group. Sensitivity of predictive species distribution models to change in grain size. **Diversity and Distributions**, v.13, p.332-340, 2007.
- GUISAN, A.; THUILLER, W. Predicting species distribution: offering more than simple habitat models. **Ecology Letters**, v.8, p.993-1009, 2005.
- HIJMANS, R. J. Very high resolution interpolated climate surfaces for global land areas. **International Journal of Climatology**, v.25, p.1965-1978, 2005.

HIJMANS, R. J. *et al.* Assessing the geographic representativeness of genebank collections: the case of Bolivian wild potatoes. **Conservation Biology**, v.14, p.1755-1765, 2000.

HOFFMANN, W.A. The effects of fire and cover on seedling establishment in a neotropical savanna. **Journal of Ecology**, v. 84, p. 383-393, 1996. INTERNATIONAL UNION FOR CONSERVATION OF NATURE – IUCN. Red list of threatened species. Versão 2010.

HORTAL, J. *et al.* Historical bias in biodiversity inventories affects the observed environmental niche of the species. **Oikos**, v.117, p.847-858, 2008.

HUETTMANN, F. Databases and Science-Based Management in the Context of wildlife and habitat: Toward a certified iso standard for objective decision-making for the global community by using the internet. **Wildlife Management**, v.69, p.466-472, 2005.

IBAMA. 2001. Management Plan for Ubajara National Park – Phase II. Ministry of the Environment, Brasília.

IBDF. 1981. Management Plan: Ubajara National Park. Brasília, Ministry of the Environment, Brasília.

IBGE. 1985. National Atlas of Brazil. IBGE, Rio de Janeiro.

IBGE – **Brazilian Institute of Geography and Statistics**. Territorial Meshes. 2019b. Available at <<https://mapas.ibge.gov.br/bases-e-referenciais/basescartograficas/malhas-digitais>>. Accessed on: April 6, 2023.

IBGE – **Brazilian Institute of Geography and Statistics**. Territorial Meshes. 2019a. Available at <<https://mapas.ibge.gov.br/bases-e-referenciais/basescartograficas/malhas-digitais>>. Accessed on: April 6, 2023.

IBGE – **Brazilian Institute of Geography and Statistics**. Biomes and Coastal-Marine System of Brazil: compatible with the scale 1:250,000. Rio de Janeiro, Coordination of Natural Resources and Environmental Studies. 168 p. (Methodological reports, v. 45, 2019).

IBAMA. 2009. Catimbau National Park. Available at: <http://www.ibama.gov.br/siucweb/mostraUc.php?seqUc=1438>. Accessed on: March 20, 2018.

BRAZILIAN INSTITUTE OF GEOGRAPHY AND STATISTICS. **Biomes and Coastal-Marine System of Brazil**. 2019. Available at: <<https://www.ibge.gov.br/apps/biomas/#/home>>. Accessed on: July 29, 2022.

BRAZILIAN INSTITUTE OF GEOGRAPHY AND STATISTICS (IBGE). Technical Manuals in Geosciences. n. 1. Technical Manual of Brazilian Vegetation. 2nd ed. Rio de Janeiro: IBGE, 2012. Available at: <https://biblioteca.ibge.gov.br/visualizacao/livros/liv63011.pdf>. Accessed on: March 10, 2019.

BRAZILIAN INSTITUTE OF GEOGRAPHY AND STATISTICS (IBGE). Ceará. 2019. Available at: <https://cidades.ibge.gov.br/brasil/ce/panorama>. Accessed on: May 5, 2019.

INSTITUTE OF RESEARCH AND ECONOMIC STRATEGY OF CEARÁ (IPECE). Ceará in numbers. 2016. Available at: http://www2.ipece.ce.gov.br/publicacoes/ceara_em_numeros/2016/territorial/index.htm. Accessed on: February 25, 2019.

IVANOV, Marlete Moreira Mendes; LEMOS, Jesus Rodrigues (eds.). Conservation Units of the State of Piauí: Volume 2. Teresina: **IFPI**, 2022.

IVANAUSKAS, N.M.; MONTEIRO, R.; RODRIGUES, R.R. Floristic similarity among Atlantic forest areas in the state of São Paulo. **Brazilian Journal of Ecology**, v. 1, n. 2, p. 71-81, 2000.

JOLY, A.B. Understanding Brazilian vegetation. São Paulo: University of São Paulo Publisher: Polígono, 1970. 165 p.

KIM, A.C. Lianas of the Atlantic Forest in the State of São Paulo. 1996. 221 f. Dissertation (Master's in Plant Biology) – Institute of Biology, State University of Campinas, Campinas.

KENDAL, D.; WILLIAMS, N. S. G.; WILLIAMS, K. J. H. A cultivated environment: exploring the global distribution of plants in gardens, parks and streetscapes. **Urban Ecosystems**, v.15, n.3, p.637-652, 2012.

KÖPPEN, W. Climatology: with a study of the climates of the Earth. México: Fondo de Cultura Económica, 478 p. 1948.

LACERDA, Francinete Francis et al. Climate trends in the semi-arid region facing global climate change perspectives; the case of Araripina, Pernambuco. **Revista do Departamento de Geografia**, v. 31, p. 132-141, 2016.

LEAL, I. R.; TABARELLI, Marcelo; SILVA, JMC da. Ecology and conservation of the caatinga: an introduction to the challenge. **Ecology and conservation of the caatinga**, p. 13-18, 2003.

LIMA-RIBEIRO, M. S., NOGUÉS-BRAVO, D., Terribile, L. C., Batra, P. & Diniz-Filho, J. A. F. Climate and humans set the place and time of proboscidean extinction in late Quaternary South America. **Paleogeogr. Paleoclimatol.** Paleoecol. 392, 546-556, 2013.

LIMA, L.C.; MORAIS, J.O.; SOUZA, M.J.N. Territorial compartmentalization and regional management of Ceará. Fortaleza: **FUNECE**, 268p. 2000.

LOIOLA, Maria Iracema Bezerra et al. Diversity of angiosperms in Ceará. **Herbário Prisco Bezerra**, v. 80, p. 1-260, 2020.

LOIOLA, Maria Iracema Bezerra et al. Diversity of angiosperms in Ceará: Herbário Prisco Bezerra-80 years of history. **Sobral: Edições UVA**, 2020.

LOPES, Sérgio de Faria et al. Seed dispersal of uruvalheira (*Platypodium elegans* VOG.) (Fabaceae) in a cerrado, Uberlândia-MG. **Revista Árvore**, v. 34, p. 807-813, 2010.

MACIEL, A.C.R. et al. Environmental quality and land use behavior in the Parnaíba Delta. **Revista Ibero-Americana de Ciências Ambientais**, v.11, n.7, p.179-187, 2020.

MACHADO, I. C. S., L. M. BARROS & E. V. S. B. SAMPAIO. Phenology of caatinga species at Serra Talhada, PE, Northeastern Brazil. **Biotropica** 29: 57-68. 1997.

MARENGO, José Antonio; CUNHA, Ana Paula Martins do Amaral; ALVES, Lincoln Muniz. The 2012-15 drought in the semiarid Northeast Brazil in historical context. **Revista Climanálise**, v.3, p. 49-54, 2016.

MARTINS-DA-SILVA, R. C. V. **Collection and Identification of Botanical Specimens**. Belém: EMBRAPA, 40 p. 2002.

MARTINS-DA-SILVA, R. C. V.; FERREIRA, G. C. **The Herbarium Collection IAN 1: Meliaceae and Ventenat**. Belém: EMBRAPA-CPATU, 58p. 1998.

MARMION, M.; PARVIAINEN, M.; LUOTO, M.; HEIKKINEN, R.K.; THUILLER, W. Evaluation of consensus methods in predictive species distribution modelling. **Diversity and Distributions**, [S. l.], v.15, n.1, p. 59–69, 2009.

Menezes, M.O.T.; Araújo, F.S.; Romero, R.E. The biological conservation system in the state of Ceará: diagnosis and recommendations. **Revista Eletrônica do Prodepa**, v. 5, n. 2, p. 7-31, 2010.

Mendes, K.; Gomes, P.; Alves, M. Floristic inventory of an ecologically sensitive area in the Northeast Brazilian Atlantic Forest. **Rodriguésia**, 61(4): 669–676, 2010.

METZ, C. E. ROC methodology in radiologic imaging. **Investigational Radiology**, v.21, p.720-733, 1986.

MISTRY, J.; BERARDI, A.; DURIGAN, G. The influence of fire regime on microscale structural variation and patchiness in Cerrado vegetation. **Rev. Inst. Flor.**, v. 22, n. 1, p. 33-49, 2010.

MITTERMEIER, R.A.; P. ROBLES-GIL; MITTERMEIER, C.G. Megadiversity: Earth's biological wealthiest nations. Mexico: **Agrupacion Sierra Madre**, 504 p. 1997.

MITTERMEIER, R.A. et al. Hotspots: Earth's biologically richest and most endangered terrestrial ecoregions. Mexico: **CEMEX**. 430 p. 1999.

MMA (Ministry of the Environment). **The Atlantic Forest Biome**. Available at: <http://www.mma.gov.br/biomas/mata-atlantica_emdesenvolvimento.html> Accessed on: 29 Nov. 2018.

MMA (Ministry of the Environment). 2014. Mapping of the vegetation cover of Brazilian biomes. Available at: <http://www.mma.gov.br/estruturas/sbf_chm_rbbio/_arquivos/mapas_cobertura_vegetal.pdf> Accessed on: 01 Nov. 2014.

MMA (Ministry of the Environment). Areas Prioritárias para a Conservação, Uso Sustentável e Repartição de Benefícios da Biodiversidade Brasileira: Atualização – Portaria MMA no. 09, de 23 de janeiro de 2007. Brasília, DF, 2007. (Série Biodiversidade).

MMA (Ministry of the Environment). Manual de Apoio Utilização das Áreas Prioritárias para a Biodiversidade da Zona Costeira e Marinha e Mata Atlântica. Brasília, DF 102 p. 2022.

MMA. Brazilian biodiversity: assessment and identification of priority areas for conservation, sustainable use and sharing of benefits of Brazilian biodiversity. MMA, Brasília, Brazil, 404pp, 2003.

MONTEIRO, P. S. D. Proposal to improve distribution maps of taxa with scarce data. Dissertation (Master's in Biodiversity and Evolution) - Museu Paraense Emílio Goeldi, Belém, 73 p. 2017.

MONTEIRO, C.A. Climatic dynamics and rainfall in the State of São Paulo – a geographical study in the form of an atlas. São Paulo: FAPESP: Institute of Geography: University of São Paulo. 129 p. 1973.

MOREIRA, H. M. Phenological studies in a remnant of caatinga in the hinterland of Paraíba. Bachelor's thesis in Biological Sciences, Federal University of Paraíba, João Pessoa. 1996.

MORELLATO, L. P. C., R. R. RODRIGUES, H. F. LEITÃO-FILHO & C. A. JOLY. Comparative study of the phenology of tree species in montane forests and semideciduous mesophytic forests in the Serra do Japi, Jundiaí, São Paulo. **Brazilian Journal of Botany** 12: 85-98. 1989.

MORELLATO, L. P. C. & H. F. LEITÃO-FILHO. Phenological strategies of tree species in mesophytic forests in the Serra do Japi, Jundiaí, São Paulo. **Brazilian Journal of Biology** 50: 163-173. 1990.

MORELLATO, L. P. C. & H. F. LEITÃO-FILHO. Fruit production and dispersal patterns in the Serra do Japi. Pp 112-141. 1992. In: L.P.C. Morellato (ed.) **Natural History of Serra do Japi: ecology and preservation of a forest area in southeastern Brazil**. Editora da UNICAMP/FAPESP, Campinas.

MOREIRA, H. M. Phenological studies in a remnant of caatinga in the hinterland of Paraíba. Bachelor's thesis in Biological Sciences, Federal University of Paraíba, João Pessoa. 1996.

MOROZ, I.C.; CANIL, K.; ROSS, J.L.S. Environmental problems in the watershed protection areas of the São Paulo metropolitan region. Department of Geography Review, p. 35-48, 1994.

MORELLATO, L. P. C. *et al.* Phenology of Atlantic rain forest trees: a comparative study. **Biotropica**, v.32, p.811-823, 2000.

MORELLATO *et al.*, - STRADIC, S.L.; MORELLATO, L.P.C.; NEVES, F.S.; OLIVEIRA, R.S.; SCHAEFER, C.E.; VIANA, P.L.; LAMBERS, H. Ecology and evolution of plant diversity in the endangered campo rupestre: a neglected conservation priority. **Plant and Soil**, v.403, p.129-152, 2016.

MORO, M.F.; LUGHADHA, E.N.; Filer, D.L.; Araújo, F.S. & Martins, F.R. A catalogue of the vascular plants of the Caatinga Phytogeographical Domain: a synthesis of floristic and phytosociological surveys. *Phytotaxa* 160: 1-118. 2014.

MORRONE, J.J. Beyond binary oppositions. *Cladistics*, 9: 437–438. 1993.

MORRONE, J.J. **Biogeographical homology**: spatial coordinates of life. Cuadernos del Instituto de Biología 37, Instituto de Biología, UNAM, Mexico City, 2004.

MYERS, N.; MITTERMEIER, R. A.; MITTERMEIER, C. G.; Da FONSECA, G. A. ;KENT, J. Biodiversity hotspots for conservation priorities. *Nature*, 403(6772), 853. 2000.

NAIMI, B. *et al.* Where is positional uncertainty a problem for species distribution modelling? **Ecography**, [S. l.], v.37, n.2, p.191–203, 2014.

NAIMI, B.; ARAÚJO, M. B. sdm: a reproducible and extensible R platform for species distribution modelling. **Ecography**, [S. l.], v.39, n.4, p.368–375, 2016.

NAKAJIMA, J. *et al.* Asteraceae. In: FORZZA, R.C. *et al.* (Coord.). List of species of the flora of Brazil 2010. Available at: . Accessed on June 8, 2023.

NUSBAUMER, L., M. R. V. Barbosa, W. W. Thomas, M. V. Alves, P.-A. Loizeau & R. Spichiger. Flora and vegetation of the Pedra Talhada Biological Reserve. In: Studer, A., L. Nusbaumer & R. Spichiger (Eds.). Biodiversity of the Pedra Talhada Biological Reserve (Alagoas, Pernambuco - Brazil). **Boissiera** 68: 59-121., 2015.

OLIVEIRA I.S.S. (2008) Study of environmental impacts as a subsidy for the planning of trails in the Serra de Itabaiana National Park, SE. **Boletim Goiano de Geografia**, 28(1): 115–126.

OLIVEIRA-FILHO, A. T. & J. A. Ratter. 2002. Vegetation Physiognomies and Woody Flora of the Cerrado Biome. In: Oliveira P. S. & J. A. Ratter (eds.). **The Cerrados of Brazil: Ecology and**

Natural History of a Neotropical Savanna: 91120. Columbia University Press, New-York.

OLIVEIRA-FILHO, A. T. & M. A. L. Fontes. Patterns of floristic differentiation among Atlantic forests in southeastern Brazil and the influence of climate. **Biotropica** 32: 793-810. 2000.

OLIVEIRA-FILHO, A. T., J. A. Jarenkow & M. J. Nogueira Rodal. Floristic relationships of seasonally dry forests of eastern South America based on tree species distribution patterns. In: Pennington, R. T. & J. A. Ratter (eds.). **Neotropical savannas and seasonally dry forests: plant diversity, biogeography and conservation**: 159–192. 2006. CRC Press Taylor & Francis Group, Boca Raton, London, New York.

OLIVEIRA-FILHO, A. T. NeoTropTree, Arboreal Flora of the Neotropical Region: A database involving biogeography, diversity, and conservation. Universidade Federal de Minas Gerais. 2014.

OLIVEIRA, J.B. Soils of the State of São Paulo: Description of classes recorded in the soil map. Instituto Agrônomo de Campinas (SP). Campinas: Instituto Agrônomo, 108 p. 1999. (Boletim Científico IAC, v. 45). OLIVEIRA, R.J. Variation in floristic composition and alpha

diversity of Atlantic forests in the State of São Paulo. 2006. 144 f. Thesis (Ph.D. in Plant Biology) – Universidade Estadual de Campinas, Campinas.

OLIVEIRA, Z. L., R. C. B. Santos-Junior, A. L. P. Feliciano, L. C. Marangon & A. J. E. de Carvalho. 2001. Floristic and phytosociological survey of a section of Atlantic Forest at the Nisia Floresta Experimental Forest Station, RN. **Revista Brasil Florestal** 20: 22-29.

OLSON, D. M., Dinerstein, E., Wikramanayake, E. D., Burgess, N. D., Powell, G. V. N., Underwood, E. C., D'Amico, J. A., Itoua, I., Strand, H. E., Morrison, J. C., Loucks, C. J., Allnutt, T. F., Ricketts, T. H., Kura, Y., Lamoreux, J. F., Wettengel, W. W., Hedao, P., Kassem, K. R. 2001. Terrestrial ecoregions of the world: A new map of life on Earth. *Bioscience* 51(11):933-938. 2004.

PAGLIA, A. P.; REZENDE, D. T.; KOCH, I.; KORTZ, A. R.; DONATTI, C. Species distribution models in biodiversity conservation strategies and ecosystem-based adaptation to climate change. **Natureza & Conservação**, Curitiba, v.10, n.2, p.231-234, 2012.

PANERO, J. L.; FUNK, V. A. The value of sampling anomalous taxa in phylogenetic studies: Major clades of the Asteraceae revealed. **Molecular Phylogenetics and Evolution**, v.47, p.757-782, 2008.

PAULA, B. T. de, Silva, F. C. da, & Faria, E. R. de. Public policies for sustainable tourism: The case of Armação dos Búzios – RJ. **Revista Turismo Em Análise**, 31(2), 316-338. 2020.

BRAZILIAN CONSERVATION UNITS PANEL. [s.l.], 2019. Available at: <https://app.powerbi.com/view?r=eyJrIjoiMDNmZTA5Y2ItNmFkMy00Njk2LWI4YjYtZDZJINzFkOGM5NWQ4IiwidCI6IjJmY2ZmE5LTNmOTMtNGJiMS05ODMwLTYzNDY3NTJmMDNlNCIsImMiOiJF9> Accessed on: September 10, 2023.

PEARSON, R.G.; DAWSON, T.P. Predicting the impacts of climate change on the distribution of species: are bioclimate envelope models useful? **Global Ecology and Biogeography**, v.12, p.361-371, 2003.

PEARSON, R. G. Climate change and the migration capacity of species. **Trends in Ecology and Evolution**, v.21, n.3, p.111–113, 2006.

PEARSON, R. G. Species' Distribution Modeling for Conservation Educators and Practitioners. **American Museum of Natural History**, v.1, p.1–50, 2007.

PETERSON, A. T. Ecological niches and geographic distributions. **Monographs in Population Biology**, Princeton University Press, Princeton, NJ, 2011.

PETERSON, A. T. *et al.* Native-range ecology and invasive potential of *Cricetomys* in North America. **Journal of Mammalogy**, v.87, p.427-432, 2006a.

PETERSON, A. T. *et al.* Geographic potential for outbreaks of Marburg hemorrhagic fever. **American Journal of Tropical Medicine and Hygiene**, v.75, p.9-15, 2006b.

PIELOU CE. 1979 . Biogeografia. Nova Iorque: John Wiley and Sons.

PHILLIPS, S. J.; ANDERSON, R. P.; SCHAPIRE, R. E. Maximum entropy modeling of species geographic distributions. **Ecological Modelling**, [S. l.], v.190, n.3–4, p.231–259, 2006.

PNUD – **United Nations Development Programme**. Brazil, 2020. Available at: <https://www.br.undp.org/content/brazil/pt/home/countryinfo.html>. Accessed on: June 21, 2023.

QGIS Development Team, 2018. **QGIS Geographic Information System**. Open Source Geospatial Foundation Project. Disponível em: <http://qgis.osgeo.org>. Acessado em: 20 set. 2023.

QUEIROZ, L. P. . The Brazilian Caatinga: Phytogeographical patterns inferred from distribution data of the Leguminosae. *In*: Pennington, R. T. & J. A. Ratter (eds.). **Neotropical savanas and seasonally dry forests: plant diversity, biogeography and conservation**: 121–157. CRC Press Taylor & Francis Group, Boca Raton, London, New York. pp. 113-149. 2006.

RATTER, J.A. et al. Analysis of the floristic composition of the Brazilian cerrado vegetation II: comparison of the wood vegetation of 98 areas. **Edinburg Journal of Botany**, v. 53, n. 153-180, 1996.

RAMOS, V. S., G. Durigan, G. A. D. C. Franco, M. F. de Siqueira & R. R. Rodrigues. **Trees of the seasonal semideciduous forest: species identification guide** : 1-312. 2007. Editora da Universidade de São Paulo, São Paulo.

RICKETTS, T.H., E. Dinerstein, D.M. Olson, C. Loucks. Who's where in North America? Patterns of species richness and the utility of indicator taxa for conservation. **Bioscience** 49(5):369-381. 1999.

RIBEIRO, J.F. & WALTER, B.M.T. The main phytophysionomies of the Cerrado Biome. In Cerrado: ecology and flora (S.M. Sano, S.P. Almeida & J.F. Ribeiro, eds.). **Embrapa Cerrados**, Planaltina. p.151-212. 2008.

RIBEIRO, M. C. et al. The Brazilian Atlantic Forest: How much is left, and how is the remaining forest distributed? Implications for conservation. **Biological Conservation**, v. 142, n. 6, p. 1141–1153, jun. 2009.

RIBEIRO, M. C. et al. Strategies for biodiversity conservation in the Atlantic Forest incorporating multiple initiatives and scales. p. 137-165, 2013.

R CORE TEAM. R: A language and environment for statistical computing. **R Foundation for Statistical Computing**, Vienna, Austria, 2023 Disponível em: <https://www.R-project.org/>.

RESOLUTION SMA 20, March 9, 1998. Publishes the preliminary list of endangered plant species in the State of São Paulo. Official Gazette of the State of São Paulo, Executive Branch, v. 108, n. 46, Mar. 10, 1998. Section I, pp. 23-25.

RESOLUTION SMA 48, September 21, 2004. Publishes the list of endangered flora species in the State of São Paulo. Official Gazette of the State of São Paulo, Executive Branch, v. 114, n. 179, Sep. 22, 2004. Section I, pp. 26-33.

RODAL, M. J. N., M. R. V. Barbosa & W. W. Thomas. Do the seasonal forests in northeastern Brazil represent a single floristic unit? **Brazilian Journal of Biology** 68: 467-475. 2008.

RODAL, M.J.N., ANDRADE, K.V.S., SALES, M.F. & GOMES, A.P.S. Phytosociology of the woody component of a vegetational refuge in the municipality of Buique, Pernambuco. **Rev. Bras. Biol.** 58 (3): 517-526. 1998.

RODAL, M. J. N. & A. L. MELO. Preliminary survey of woody species in the caatinga of Pernambuco. In: Proceedings of the 1st Workshop General Program Plants of the Northeast. **Royal Botanic Gardens**, Kew, pp. 53-62., 1999.

ROUCOU, P.; ARAGÃO, J. O. R.; Harzallah, A.; Fontaine, B.; Janicot, S. Vertical motion, changes to Northeast Brazil rainfall variability: A GCM simulation. **International Journal of Climatology**, Chichester, v.16, n.1, p.879-891, 1996.

ROQUE, N.; NAKAJIMA, J. N. Two new species of *Richterago* Kuntze emend. Roque (Asteraceae, Mutisieae) from Minas Gerais and Goiás, Brazil. **Kew Bulletin**, v.56, p.697-703, 2001.

ROQUE, N.; BAUTISTA, H. P. **Asteraceae: Characterization and Floral Morphology**. EDUFBA, Salvador, 2008. 69p.

ROQUE, N.; FUNK, V. A. Morphological characters add support for some members of the basal grade of Asteraceae. **Botanical Journal of Linnean Society**, v.171, p.568–586, 2013.

Roque, N., Sancho, G. 2020. *Moquiniastrium* in Flora do Brasil 2020. Jardim Botânico do Rio de Janeiro.

ROOT, T. L. *et al.* Fingerprints of global warming on wild animals and plants. **Nature**, v.421, p.57-60, 2003.

RUGGIERO, P. G. C., M. A. Batalha, V. R. Pivello & S. T. Meirelles. Soil-vegetation relationships in cerrado (Brazilian savanna) and semideciduous forest, Southeastern Brazil. **Plant Ecology** 160: 1-16. 2002.

SÁBER, A. N. A. & L. C. Marigo. Brazilian ecosystems: 1-299. 2009. **Metalivros**, São Paulo.

SALES, M.F., MAYO, S.J. & RODAL, S.J. **Vascular plants of the montane forests of Pernambuco: a checklist of threatened flora of the Brejos de Altitude Pernambuco, Brazil**. Recife, Universidade Federal Rural de Pernambuco, 130 pp. 1998.

SAMPAIO, E. V. S. B. Overview of the Brazilian caatinga. *In*: Bullock, S. H., H. A. Mooney & E. Medina (eds.). **Seasonally Dry Tropical Forests**: 35-63. Cambridge University Press. 1995

SANCHO, G.; ROQUE, N. *Moquiniastrium* in **Flora e Funga do Brasil**. Jardim Botânico do Rio de Janeiro. Disponível em: <https://floradobrasil.jbrj.gov.br/FB129778>. Acesso em: 13 jun. 2023.

SANCHO, G.; FUNK, V. A; ROQUE, N. *Moquiniastrium* (Gochnatieae, Asteraceae): disentangling the paraphyletic *Gochnatia*. **Phytotaxa**, v.147, p.26-34, 2013.

SANCHO, S. M., S. P. Almeida & J. F. Ribeiro. 2008. **Cerrado: ecology and flora**: 1-1279 EMBRAPA Informação Tecnológica : Planaltina, DF: EMBRAPA Cerrados, Brasília.

SANCHO, G.; FREIRE, S. Gochnatieae (Gochnatioideae) and Hyalideae (Wunderlichioideae p.p). *In*: Systematic evolution and biogeography of the Compositae, Funk Vicky; STUESSY, Adryan. **BAYER**, Taylor, Viena, Austria, IATP, 2009.

SANCHO, G. Revisión y filogenia de la sección Moquiniastrun Cabrera del género Gochnattia Kunth (Asteraceae, Mutisieae). **Fontqueria**, v.54, p.61-122, 2000.

SANCHO, G.; Roque, N. Moquiniastrum in **Flora and Funga of Brazil**. Rio de Janeiro Botanical Garden. Available at: <https://floradobrasil.jbrj.gov.br/FB130865>. Accessed on: 14 Feb. 2024.

SÃO PAULO (State). Decree No. 36,859, of 5 June 1993. Establishes the Juquery State Park and related measures. Official Gazette of the State of São Paulo, Executive Branch, v. 103, n. 106, 8 June 1993. Section I, pp. 2-3.

SCUDELLER, V.V. Phytogeographical analysis of the Atlantic Forest – Brazil. 2002. 204 p. Thesis (PhD in Plant Biology) – State University of Campinas, Campinas.

SEPLAN, Atlas of Sergipe. Aracaju: Federal University of Sergipe, State Planning Secretariat of Sergipe. 95 p., 1979.

SEIBERT, P. **Guide to South America**: landscapes and vegetation: 1-271. 1998. Eugen Ulmer Editions, Paris.

SCHRIRE, BRIAN D. et al. Global distribution patterns of the Leguminosae: insights from recent phylogenies. In: Plant diversity and complexity patterns: local, regional and global dimensions. Proceedings of an International Symposium held at the Royal Danish Academy of Sciences and Letters in Copenhagen, Denmark, 25-28 May, 2003. **Det Kongelige Danske Videnskabernes Selskab**, p. 375-422. 2005.

SANCHO, G.; ROQUE, N. Moquiniastrum in Flora e Funga do Brasil. **Jardim Botânico do Rio de Janeiro**. Available at: <https://floradobrasil.jbrj.gov.br/FB129778>. Acesso em: 13 jun. 2023.

SIQUEIRA, L. P. **Monitoring of restored areas in the interior of São Paulo State, Brazil**. 2002. Dissertation (Master's in Forest Resources), Luiz de Queiroz College of Agriculture, University of São Paulo, Piracicaba-SP, 2002.

SILLERO, N., ARENAS-CASTRO, S., ENRIQUEZ-URZELAI, U., C. Gomes Vale, D. Sousa Guedes, F. Martínez-Freiría, R. Real, A. Márcia Barbosa. Want to model a species niche? A step-by-step guideline on correlative ecological niche modelling. **Ecol. Model.**, 456 2021, 10.1016/j.ecolmodel.2021.109671.

SILVA, J. M. C., M. TABARELLI, M. T. FONSECA & L. V. LINS. **Biodiversity of the caatinga: priority areas and actions for conservation**: 1-382. 2004. Brasília: Ministry of the Environment, Recife: Federal University of Pernambuco.

SILVA A.C.C., Oliveira E.V.S., Alves M., Farias M.C.V., Mota A.C., Souza C.A.S. & Prata A.P.N. Updated list of vascular flora of Serra de Itabaiana National Park (PARNA), Sergipe,

Brazil. **Pesquisa e Ensino em Ciências Exatas e da Natureza**, 3(1):40–67. 2019
<http://dx.doi.org/10.29215/pecen.v3i1.1148>

SIGEP - Brazilian Commission of Geological and Paleobiological Sites. 2010. Available at:
<http://www.unb.br/ig/sigep>. Accessed on December 20, 2023.

SPECIESLINK. *speciesLink*. Available at: <https://specieslink.net/>. Accessed on: July 5, 2023.

SPLINK. Data and Tools. Available at: www.splink.cria.org.br. Accessed on: January 7, 2023.

SIQUEIRA, M. F.; DURIGAN, G. Modeling the geographic distribution of woody species in cerrado, São Paulo State, Brazil. **Revista Brasileira de Botânica**, v.30, p.233-249, 2007.

SOBRAL I.S., Santana R.K.O., Gomes L.J., Costa M., Ribeiro G.T. & Santos J.R. Evaluation of environmental impacts in Serra de Itabaiana National Park - SE. **Caminhos de Geografia**, 8(24): 102–110; 2007.

SOBERÓN, J. Niche and area of distribution modeling: a population ecology perspective. **Ecography**, v.33, p.159-167, 2010.

SOUZA, L. S. de; et al. Environmental perception of the caatinga biome in the school context. **Revista Iberoamericana de Educación**, 2017.

SOUZA, N. S.; PAIVA, C. C. Water in the Semi-arid: Divergent Discourses and Practices. ComSertões: **Journal of communication and culture in the semi-arid**, n. 5, Jul/Dec, 2017 - Juazeiro: UNEB/DCH, 2017.

STUDER, A. Ecological study of the forest complex of Serra das Guaribas and Serra do Cavaleiro. **Request for the Safeguarding of this Forest**: 1-61 1985. Quebrangulo, Alagoas, Brazil.

TABARELLI, M., L. P. PINTO, J. M. C. Silva, M. Hirota & L. Bede. Challenges and opportunities for Biodiversity conservation in the Brazilian Atlantic forest. **Conservation Biology** 19: 695-700. 2005

TABARELLI, MARCELO et al. Prospects for biodiversity conservation in the Atlantic Forest: lessons from aging human-modified landscapes. **Biological Conservation**, v.143, n.10, p.2328-2340, 2010.

TABARELLI, M. et al. Caatinga: legacy, trajectory and challenges towards sustainability. **Cienc. Cult.** São Paulo, v. 70, n. 4, p. 25-29, Oct. 2018.

TABARELLI, M., J. M. C. SILVA, A. VICENTE & A. M. SANTOS. Analysis of representativeness of direct and indirect use conservation units in the Caatinga: preliminary analysis. Pp. 13 2000. in: SILVA, J. M. C. & M. TABARELLI (coords.) **Workshop Evaluation and identification of priority actions for conservation, sustainable use and benefit sharing of biodiversity in the Caatinga biome**. Petrolina, Pernambuco, Brazil. www.biodiversitas.org.br/caatinga.

TABARELLI, M.; SILVA, J. M. C. da. Áreas e ações prioritárias para a conservação da biodiversidade da Caatinga. **Ecologia e conservação da Caatinga**, p. 777- 796, 2003.

TABARELLI, M.; SILVA, J. M. C. da. Priority areas and actions for biodiversity conservation in the Caatinga. **Ecology and Conservation of the Caatinga**, p. 777-796, 2003.

TABARELLI, M.; MANTOVANI, W. Natural gaps and species richness of pioneer species in a montane Atlantic forest. **Revista Brasileira Biologia.**, v. 59, n. 2, p. 251-261, 1999.

TEIXEIRA LP, Lughadha EM, Silva, MVC, Moro MF. How much of the Caatinga is legally protected? Analysis of temporal and geographical coverage of protected areas in the Brazilian semi-arid. **Acta Botânica Brasileira** 2021; 35(3): 473-485.

TSCHARNER, T., G. P. Duda, V. P. Oliveira, C. M. S. Silva, L. Nusbaumer & A. F. Silva Filho. Abiotic parameters of the Pedra Talhada Biological Reserve. In: Studer, A., L. Nusbaumer & R. Spichiger (Eds.). Biodiversity of the Pedra Talhada Biological Reserve (Alagoas, Pernambuco - Brazil). **Boissiera** 68: 39-57. 2015.

UDVARDY MDF. A classification of the world biogeographical provinces Morges (Switzerland): International Union for Conservation of Nature and Natural Resources. Occasional Paper of IUCN No. 18. 1975.

VELLOSO, A. L.; SAMPAIO, E. V. S. B.; PAREYN, F. G. C. Proposed ecoregions for the Caatinga biome: results from the ecoregional planning seminar of the Caatinga / Aldeia-PE November 28-30, 2001. **Associação Plantas do Nordeste, Instituto de Conservação Ambiental - The Nature Conservancy do Brasil**, Recife, 2002. 76 p.

VICENTE, A. et al. Botanical survey. In: CARVALHO, C. M.; VILAR J. C. (coords). **Serra de Itabaiana National Park - Biota Survey**. Aracaju: Ibama, General and Experimental Biology-UFS, p.15-37, 2005.

VICENTE A., RIBEIRO A.S., SANTOS E.A. & Franco C.R.P. Botanical Survey (p. 15–37). In: Carvalho C.M. & Vilar J.C. (Coords). Serra de Itabaiana National Park - Biota Survey. Aracaju: Ibama, **General and Experimental Biology-UFS**. 131 p.; 2005.

WANG, S. et al. Response of spatial vegetation distribution in China to climate changes since the Last Glacial Maximum (LGM). **PLoS ONE**, v.12, n.4, p.1-18, 2017.

WARREN, D. L. et al. Incorporating model complexity and spatial sampling bias into ecological niche models of climate change risks faced by 90 California vertebrate species of concern. **Diversity and distributions**, v.20, n.3, p.334-343, 2014.

WISZ, M. S. et al. The role of biotic interactions in shaping distributions and realized assemblages of species: Implications for species distribution modelling. **Biological Reviews**, v.88, n.1, p.15–30, 2013.

WORLDCLIM, **WorldClim**: global climate data: free climate data for ecological modeling and GIS. Museum of Vertebrate Zoology, University of California, Berkeley, USA. Accessed: Aug 2, 2022. Athiê-Souza, S. M., Melo, J. I. M., Silva, L. P., Santos, L. L., Santos, J. S., Oliveira, L. S.

D., & Sales, M. F. Phanerogamic flora of the Catimbau National Park, Pernambuco, Brazil. *Biota Neotropica*, v.19. p.37, 2019.

WWF – World Wide Fund for Nature. **Wildfinder Database**. 2023. Available at: <https://www.worldwildlife.org/publications/wildfinder-database>.

ZANELLA, M. E. (2014). Considerations on the climate and water resources of the Northeastern semi-arid region. *Caderno Prudentino De Geografia*, 1(36), 126–142. Retrieved from <https://revista.fct.unesp.br/index.php/cpg/article/view>